

1 Original Article

2
3 **Phenolic concentrations of brown seaweeds and relationships to nearshore**
4 **environmental gradients in Western Australia**

5
6 Daniel H van Hees^{1,2*}, Ylva S Olsen^{1,2}, Thomas Wernberg^{1,2}, Kathryn L Van
7 Alstyne³, Gary A Kendrick^{1,2}

8
9 1. School of Biological Sciences, University of Western Australia, 35 Stirling Hwy,
10 Crawley WA 6009, Australia

11
12 2. The Oceans Institute, University of Western Australia, 35 Stirling Hwy, Crawley
13 WA 6009, Australia

14
15 3. Shannon Point Marine Laboratory, Western Washington University, 1900 Shannon
16 Point Road Anacortes WA 98221, United States

17
18
19
20 *Corresponding Author:

21 Daniel H van Hees
22 School of Biological Sciences
23 35 Stirling Highway
24 Crawley, WA 6009 Australia
25 email: daniel.vanhees@research.uwa.edu.au
26 phone: +61 422 239 865
27
28

29
30 Running Title: Macroalgal phenolic compounds across latitude

31
32
33
34
35
36
37
38

39 ABSTRACT

40 Phenolic compounds are found in all brown macroalgae and function as cell wall
41 structure, UV protection and as herbivore deterrents. The concentrations of phenolic
42 compounds vary among taxa and between temperate and tropical ecosystems.
43 Australasia has high concentrations of soluble phenolics compared to other regions.
44 Presently, relationships between phenolic concentrations and environmental gradients
45 are unclear. The purpose of this study was to determine the soluble phenolic
46 concentrations of brown seaweeds along temperate and tropical ecosystems of the
47 Western Australia coastline. We tested the hypothesis that phenolic concentrations are
48 related to local and broad-scale abiotic environmental gradients. Strong environmental
49 gradients of coastal Western Australia provided the opportunity to characterize
50 phenolic compounds across one large gradient. Phenolic concentrations of brown
51 seaweeds at seven study locations varied across latitude with higher concentrations
52 found at higher latitudes and were comparable to seaweeds from similar latitudes in
53 Australia. This trend coincided with a negative relationship between
54 photosynthetically active radiation and phenolic compounds, and a positive
55 relationship with salinity. We also found phenolic concentrations were positively
56 related to salinity in tropical Shark Bay but this was dependent on species.
57 Environmental conditions are important in regulating concentrations of phenolic
58 compounds. Multiple factors influence the concentrations of macroalgal phenolic
59 compounds creating unique distributions among geographical regions. This study
60 highlighted the importance of considering multiple factors when studying phenolic
61 ecology and suggests photosynthetically active radiation and salinity as important
62 drivers of phenolic compound distribution in Western Australia.

63

64 Keywords: latitudinal gradient, macroalgae, photosynthetically active radiation,
65 polyphenolics, salinity

66

67 **INTRODUCTION**

68 Phenolic compounds are chemically diverse primary and secondary metabolites found
69 in terrestrial and aquatic primary producers that provide cellular and ecological
70 functions through their oxidative capacity (Appel 1993; Appel et al. 2001). In marine
71 macroalgae, phenolic compounds are most common in brown seaweeds, where they
72 are composed of polymers of phloroglucinol, and often are referred to as
73 phlorotannins. Globally, phenolic concentrations of brown seaweeds range from
74 negligible to 20 % of tissue dry mass (Ragan and Glombitza 1986) with higher
75 concentrations generally found at higher latitudes (Steinberg and Paul 1990; Steinberg
76 et al. 1991). This pattern stands in contrast to other chemical compounds used in
77 marine allelopathic interactions that are generally more abundant in tropical
78 environments (Hay 1996). While globally structured by latitude, phenolic compounds
79 can occur regionally in high concentrations in low latitude areas such as the Caribbean
80 (Targett et al. 1992). Studies of phenolic compounds on regional scales provide
81 conflicting relationships to latitude and associated environmental conditions
82 (Steinberg 1989; Van Alstyne et al. 1999; Le Lann et al. 2012; Tanniou et al. 2014).

83 Phenolic concentrations differ across spatial scales ranging from a few meters to
84 1000's of kilometers (Steinberg 1989; Pavia and Åberg 1996; Van Alstyne et al.
85 1999; Le Lann et al. 2012) as well as among closely (Van Alstyne et al. 1999; Connan
86 et al. 2004) and distantly related taxa (Steinberg 1989).

87

88 Although ubiquitous in brown seaweeds, the distribution of, and factors that relate to,
89 phenolic compounds are not well understood at regional and global scales. Phenolic
90 compounds primarily strengthen cell walls (Schoenwaelder and Clayton 1999) and
91 function in wound healing and repair (Lüder and Clayton 2004). They are also known
92 to protect against abiotic stressors like ultraviolet radiation (Swanson and Druehl
93 2002; Jormalainen and Honkanen 2004) and salinity by acting as oxidant scavenging
94 molecules (Pedersen 1984; Ragan and Glombitza 1986). In addition to cellular
95 functions, phenolic compounds act as allelopathic compounds and defend seaweeds
96 against some herbivores (Amsler and Fairhead 2006) as well as fungal and bacterial
97 infections (Plouguerné et al. 2012; Tanniou et al. 2014). Phenolic concentrations in
98 brown seaweeds are generally higher in Australasia than in the northern hemisphere
99 (Steinberg 1989; Steinberg and Van Altena 1992; Van Alstyne et al 1999; Yates and
100 Peckol 1993). This difference is possibly a result of differences in herbivory pressures
101 between the two regions (Estes et al 1988; Steinberg et al. 1995). Because of this, we
102 examined seaweeds of Western Australia to see if they follow the typical trend of
103 higher concentrations at higher latitudes and higher concentrations than the northern
104 hemisphere.

105

106 Western Australia is a global hotspot for temperate marine macroalgal diversity
107 (Wernberg et al. 2011; Bennett et al. 2016) but the distribution of phenolic
108 compounds in this region is currently unknown. The Leeuwin Current is the dominant
109 ocean current along nearshore Western Australia. It originates in the Indonesian
110 through-flow and continues around the southern coast of Australia across the Great
111 Australian Bight (Wernberg et al. 2013) carrying warm tropical waters south to higher
112 latitudes. This creates strong abiotic gradients against which many ecological

113 gradients have been studied (McGowran et al. 1997; Kendrick et al. 2009; Wernberg
114 et al. 2013). Along these gradients, herbivore activity is similar to global patterns
115 (Vermeij 1978; Gaines and Lubchenco 1982), with higher grazing rates at lower
116 latitudes (Vanderklift and Kendrick 2004; Poore et al. 2012). Additionally, localized
117 abiotic gradients exist. For example, a salinity gradient in tropical Shark Bay (Figure
118 1) ranges from marine to salinities higher than 60 psu at the southern terminus
119 (Walker 1985). Additionally, tidal exchanges within tropical Cygnet Bay can be as
120 large as 10 m resulting in an extremely variable intertidal environment. Local
121 gradients such as these, along with the geographic scope of Western Australia,
122 represent a unique opportunity to advance the understanding of marine phenolic
123 compound distribution on continental scales.

124

125 In this study, we quantified phenolic compounds of dominant brown seaweeds in
126 Western Australia and investigated potential relationships of phenolic concentrations
127 to environmental conditions among taxa across broad and fine spatial scales
128 throughout coastal Western Australia. We specifically aimed to answer the following
129 questions: 1) What variability of soluble phenolic concentrations exists among
130 macroalgal taxa at different spatial scales? 2) How do phenolic concentrations of
131 brown seaweeds from nearshore Western Australia compare to other regions? 3) Do
132 phenolic concentrations of seaweeds correlate with local environmental gradients?
133 These questions were investigated by sampling brown seaweeds from seven
134 geographically distinct nearshore study locations of Western Australia and comparing
135 those findings to environmental conditions gathered both *in situ* and from satellite
136 imagery.

137

138 **METHODS**

139 **Study Locations**

140 Brown marine seaweeds (Class Phaeophyceae) were collected from high intertidal to
141 12 m depth at seven nearshore study locations in Western Australia (Figure 1) from
142 March to May 2015. One to seven sites were sampled (Table 1) within each location
143 to cover known within-location gradients. The study locations represent temperate (3
144 locations) and tropical (4 locations) systems in Western Australia and span 17 degrees
145 of latitude (Table 1). Thirty-one species of seaweed were haphazardly collected
146 during this study. Species continuity of collections was maintained when possible, but
147 this was infrequent and mostly different algal species were collected at study locations
148 separated by 150 -1000 km (Figure 1). We collected *Hormophysa cuneiformis* at three
149 locations, *Padina sp.* at three locations, and *Sargassum ligulatum* and *Sirophysalis*
150 *trinodis* at five locations (Table 2). Seaweeds collected during this study represented
151 the dominant habitat-forming brown taxa present at each study location. The
152 taxonomic families collected were: Dictyotaceae, Lessionaceae, Sargassaceae, and
153 Sporochneaceae. Both Dictyotaceae and Sargassaceae were distributed throughout
154 sampling locations, comprising 94 % of species collections whereas Lessionaceae and
155 Sporochneaceae were represented by one species and location each.

156

157 Seaweeds were collected across local environmental gradients of depth and salinity
158 (Table 1, Figure 1). In tropical Shark Bay, three species of macroalgae; *Hormophysa*
159 *cuneiformis*, *Sargassopsis decurrens* and *S. trinodis* were collected at seven sites
160 encompassing salinities from 39.0 to 53.5 (Walker 1985) (Figure 1, Table 1).
161 *Sargassum paradoxum* and the subtidal kelp *Ecklonia radiata* were collected at three
162 sites within temperate Jurien Bay at depths ranging from 5.0 to 10.0 m. *Sargassum*

163 *paradoxum* was also collected at three sites within temperate Port Gregory ranging
164 from 1.5 - 11.0 m depth (Table 1). At the intertidal study location of tropical Cygnet
165 Bay, the tidal range is up to 10 m creating a large and dynamic intertidal zone.
166 *Hormophysa cuneiformis* and *Turbinaria gracilis* were collected at four high intertidal
167 sites within Cygnet Bay (Table 1).

168

169 **Environmental parameters**

170 Abiotic environmental conditions for each of the seven study locations across
171 nearshore Western Australia were gathered from MODIS Aqua at 4 km resolution
172 from seasonal composites averaged over the sampling period (March – May 2015)
173 (Table 3). Using QGIS we extracted data for diffuse light attenuation (at 490nm using
174 the KD2 algorithm), daytime sea surface temperature (SST), photosynthetically
175 available radiation (PAR) and chlorophyll-a concentration. Mean salinity values for
176 each study site were taken from the CSIRO Atlas of Regional Seas (CARS) as a five-
177 year average from 2003 – 2008 (Ridgway 2009).

178

179 **Tissue collection and chemical analysis**

180 Meristem tissue samples of healthy-looking brown seaweeds individuals were
181 haphazardly collected at each study location (n = 3 - 24 different plants per species)
182 and kept cool in an icebox in the field. Seaweeds were identified to the lowest
183 taxonomic level (Womersley 1987; Huisman 2000; Huisman 2015). A representative
184 of each species was lodged at the University of Western Australia Herbarium.
185 Seaweeds were cleaned of epiphytes and debris and frozen at -20 °C while in the field.
186 Frozen seaweeds were transported to the University of Western Australia no later than

187 one week after collection where, upon arrival, samples were stored at -80 °C. Tissue
188 samples were later freeze-dried prior to chemical analysis.

189

190 We analyzed soluble phenolic compounds using a modified Folin-Ciocalteu method
191 from Van Alstyne (1995). This assay measures the redox activity of a methanolic
192 extract by reacting with the available hydroxyl groups accessible to the reagents as
193 well as non-polar hydroxylated aromatic compounds that comprise < 5 % of all Folin-
194 Ciocalteu compounds (Van Alstyne 1995). Freeze-dried seaweed samples were
195 ground to a fine powder using a mixer/mill. Ten milligrams of ground tissue from
196 each individual sample were extracted in 1.0 ml 80 % methanol for 24 hours in the
197 dark. Extracts were either analyzed the following day or stored at -80 °C until
198 analysis. Forty microliters of Folin-Ciocalteu reagent were added to 100 µl of extract
199 (diluted 1:20 with Type-1 reagent grade water). The solution was then made alkaline
200 by the addition of 100 µl Na₂CO₃ after 5 minutes. Absorbance was read at 765 nm
201 using a FLUOStar microplate reader after a 30-minute incubation at 50 °C. Three
202 replicate extracts per sample were analyzed with phloroglucinol (Sigma
203 79330_FLUKA) as a standard (Van Alstyne 1995) and converted to phenolic
204 concentrations as a percent of dry mass.

205

206 **Statistical Analysis**

207 All statistical analyses were run using R Studio (R Core Team 2012). Average
208 phenolic concentration of the macroalgae at each study location was calculated as the
209 mean of all species collected at each location. Phenolic concentration data were
210 fourth-root transformed to pass the assumptions of normality and homogeneity of
211 variance with a Cochran's Test prior to analysis. To test for differences in phenolic

212 concentrations of the study location averages, we used a 2-way nested analysis of
213 variance (ANOVA) with collection site nested within study location followed by
214 Tukey's HSD post-hoc tests (Table 4). Differences in phenolic concentrations among
215 algal families and locations were tested with a 2-way ANOVA with species nested
216 within location and family (Table 4). Differences in phenolic concentrations for
217 individual species collected at multiple locations were tested with one-way ANOVAs
218 with location as the fixed factor (Table 5).

219

220 We investigated relationships between environmental conditions and phenolic
221 concentrations with a multiple linear regression analysis. We considered Cygnet Bay,
222 the only high intertidal lagoon study location, ecologically different from the other six
223 study locations for the purpose of this analysis and excluded it from the regression
224 models. We assessed the importance of predictor environmental conditions on
225 phenolic concentration with a multiple regression analysis that included: chlorophyll-
226 a, diffuse attenuation, PAR, salinity and SST. The contribution of each factor to the
227 regression model was reported with partial eta-squared (η^2) values (Table 6). We
228 removed chlorophyll-a from the model because of collinearity with attenuation (cutoff
229 r^2 value = 0.95). We also assessed the relationship between the multiple regression
230 model and individual species distributed across multiple study locations. Phenolic
231 concentrations of *H. cuneiformis*, *S. decurrens*, and *S. trinodis* were compared to
232 salinity in tropical Shark Bay. Phenolic concentrations of *E. radiata* and *S.*
233 *paradoxum* in temperate Jurien Bay and *S. paradoxum* were compared to depth.

234

235 **RESULTS**

236 Phenolic concentrations of seaweeds in this study ranged from 0.25 ± 0.03 % to 10.99
237 ± 2.63 % of dry weight (Table 2). Of the 46 species collected, 32 were low-phenolic
238 species (< 2 % DM, Steinberg 1989) and 14 were high-phenolic species (> 2 % DM,
239 Steinberg 1989). The majority of high-phenolic species were found in the temperate
240 locations. Only three species with phenolic concentrations higher than 2 % were
241 present in tropical study locations; *Hormophysa cuneiformis*, *Padina sp.* and
242 *Turbinaria gracilis* collected in Cygnet Bay (Table 2).

243

244 Average phenolic concentrations of brown seaweeds at each study location were
245 significantly different among locations (ANOVA, $F(6,45) = 16.77$, $P < 0.001$,) (Table
246 4). Lower phenolic concentrations were generally found in tropical latitudes north of
247 Shark Bay. The highest phenolic concentration of a study location, found at the
248 highest latitude, was 11 times greater than the lowest average location concentration
249 found at tropical Thevenard Island (Figure 2a) and was higher than all other study
250 locations (Tukey's HSD, $P < 0.001$). Phenolic concentrations in tropical Cygnet Bay
251 and temperate Jurien Bay were higher than those found at study locations between
252 Port Gregory and Thevenard (Tukey's HSD, $P < 0.01$).

253

254 Phenolic concentrations were significantly different among macroalgal families
255 (ANOVA, $F(3,42) = 2.86$, $P = 0.04$) (Figure 3, Table 4). Mean phenolic
256 concentrations of the Lessionaceae were significantly higher than the Sargassaceae
257 and Dictyotaceae (Tukey's HSD, $P = 0.01$) (Figure 3). Mean phenolic concentrations
258 of macroalgal families were also significantly different among study locations
259 (ANOVA, $F(3, 298) = 6.89$, $P < 0.001$) (Table 4). The phenolic concentration of the
260 Dictyotaceae was up to 8 times higher in temperate Eagle Bay than tropical Exmouth

261 Gulf while the concentration of the Sargassaceae was 12 times higher in Eagle Bay
262 than Exmouth Gulf (Figure 4).

263

264 Phenolic concentration of individual species varied among study locations. In general,
265 the highest phenolic concentrations of species were found either in tropical Cygnet
266 Bay or temperate Eagle Bay (Figure 2b). Phenolic concentrations of *Sirophysalis*
267 *trinodis* collected at five locations (Table 2), varied up to 6.5 times between locations
268 (ANOVA, $F(4,25) = 4.44$, $P = 0.008$) (Table 5) with the highest concentrations found
269 at the highest latitude location. The phenolic concentration of the tropical *Sargassum*
270 *ligulatum* was 3.4 times higher in tropical Cygnet Bay than the tropical locations
271 Exmouth Gulf or Thevenard Island (ANOVA, $F(2,16) = 30.75$, $P < 0.001$). The
272 phenolic concentration of *Lobophora variegata* was 28 times higher in temperate
273 Eagle Bay than in tropical Exmouth Gulf (ANOVA, $F(1,13) = 690.1$, $P < 0.001$). The
274 highest phenolic concentration of *H. cuneiformis*, also found in Cygnet Bay
275 (ANOVA, $F(2,38) = 33.7$, $P < 0.001$) was 3.5 times higher than Eagle Bay or tropical
276 Shark Bay.

277

278 Additionally, phenolic concentrations of some macroalgal species varied significantly
279 among sample sites within study locations. Phenolic concentrations of *H. cuneiformis*
280 (ANOVA, $F(4,19) = 4.1$, $P = 0.01$) and *S. trinodis* (ANOVA, $F(2,10) = 12.96$, $P =$
281 0.002) (Table 5) varied up to 6 times among sample sites within tropical Shark Bay
282 while phenolic concentrations of *S. decurrens* were similar among sites (Table 2).
283 Phenolic concentrations of *H. cuneiformis* (ANOVA, $F(1,10) = 6.99$, $P = 0.02$) and *T.*
284 *gracilis* (ANOVA, $F(2,12) = 4.97$, $P = 0.03$) in tropical Cygnet Bay varied by 1.4
285 times and 1.8 times between sample sites, respectively (Table 2). Phenolic

286 concentrations of *E. radiata* and *S. paradoxum* did not differ significantly among
287 collection sites in temperate Jurien Bay.

288

289 Environmental conditions during the sampling period varied across the latitudinal
290 gradient (Table 3). Photosynthetically active radiation generally decreased with
291 latitude and ranged from 26.8 Einstein m² day⁻¹ in Eagle Bay to 43.5 Einstein m² day⁻¹
292 in tropical Cygnet Bay. Average sea surface temperatures were highest at low
293 latitudes and ranged from 19.9 °C in Eagle Bay to 28.5 °C in Cygnet Bay. There was
294 no clear latitudinal pattern in diffuse attenuation coefficients, which ranged from 0.06
295 m⁻¹ in temperate Eagle Bay to 0.29 m⁻¹ at tropical Thevenard Island or in salinity,
296 which ranged from 35.1 in tropical Exmouth Gulf to 49.5 (mean value) in Shark Bay
297 (Table 3). The salinity range among sample sites in Shark Bay, however, was 35.0 –
298 53.2.

299

300 The multiple regression model found relationships between phenolic concentration
301 averages of study locations (excluding Cygnet Bay) as well as individual species
302 phenolic concentrations. A strong relationship was found between average phenolic
303 concentrations and environmental variation across study locations (multiple
304 regression; $r^2 = 0.57$, $P < 0.001$) (Table 6a). Mean phenolic concentrations of
305 macroalgae across all study locations were positively correlated to salinity ($\eta^2 = 0.15$)
306 and negatively correlated to PAR ($\eta^2 = 0.18$). Similarly, phenolic concentrations of
307 the Dictyotaceae and Sargassaceae were positively correlated to salinity ($\eta^2 = 0.14$
308 and $\eta^2 = 0.21$, respectively) and negatively with PAR ($\eta^2 = 0.11$ and $\eta^2 = 0.27$,
309 respectively) (Table 6b). While phenolic concentrations of *S. trinodis* were negatively
310 related to PAR ($\eta^2 = 0.09$) and SST ($\eta^2 = 0.10$) (Table 6), phenolic concentrations of

311 *H. cuneiformis* and *Sargassum ligulatum* were negatively related to diffuse
312 attenuation ($\eta^2 < 0.001$ and $\eta^2 < 0.001$, respectively) and positively related to SST (η^2
313 < 0.001 and $\eta^2 < 0.001$, respectively) (Table 6c).

314

315 There were also strong relationships between phenolic concentrations of individual
316 species and local abiotic gradients at individual study locations. Phenolic
317 concentrations of *S. decurrens* ($r^2 = 0.30$, $F(1,13) = 5.61$, $P = 0.03$) and *S. trinodis* (r^2
318 $= 0.29$, $F(1,15) = 6.22$, $P = 0.02$) in tropical Shark Bay were positively related to
319 salinity (Figure 5) but not in *H. cuneiformis* (Table 6c). Phenolic concentration
320 differences among sites were not related to depth in temperate Jurien Bay or Port
321 Gregory. We found no relationship between phenolic concentration and depth in *E.*
322 *radiata* in Jurien Bay or *S. paradoxum* in Port Gregory (1.5 – 11.5m depths) and
323 Jurien Bay (3 – 10 m depths) (Table 6).

324

325 **DISCUSSION**

326 In this study, we investigated latitudinal and environmental relationships to soluble
327 phenolic concentrations in a range of brown seaweeds along the coast of Western
328 Australia spanning 17 degrees of latitude. Macroalgal phenolic concentrations in
329 Western Australia were variable across multiple spatial scales and among taxa (Table
330 2). In this study macroalgal phenolic compounds were found in higher concentrations
331 at higher latitudes and in some groups were also correlated to associated gradients in
332 salinity and PAR. Seaweeds of Western Australia generally had lower phenolic
333 concentrations than seaweeds of other global areas, but displayed similar spatial
334 variability to regions of comparable size (e.g. Steinberg 1989; Van Alstyne et al.
335 1999).

336

337 Phenolic concentrations of seaweeds were highest in the temperate latitudes of
338 Western Australia. The average phenolic concentration of temperate Eagle Bay was
339 similar to sites at comparable latitudes in eastern and southern Australia (Steinberg
340 1989) (Table 2). In general, tropical seaweeds have low concentrations of phenolics
341 (Steinberg 1986; Hay and Fenical 1988; Van Alstyne and Paul 1990) and seaweeds of
342 Western Australia follow this trend (Table 2). We found phenolic concentrations were
343 three times lower in tropical study locations than temperate ones, with the exception
344 of tropical Cygnet Bay, which had phenolic concentrations comparable to those from
345 temperate Jurien Bay (Figure 2a). This is similar to seaweeds from the Caribbean Sea
346 (Targett et al. 1992) which also had anomalously high levels of phenolics for a
347 tropical region. The largest within-family differences in phenolic concentrations
348 between locations occurred in the Dictyotaceae and Sargassaceae, which were almost
349 an order of magnitude higher in temperate Eagle Bay than tropical Exmouth Gulf
350 (Figure 4). Interestingly, *Sirophysalis trinodis* contained phenolic concentrations that
351 classified it as a high phenolic species in temperate Eagle Bay and a low phenolic
352 species in tropical locations (Table 2). Species found across temperate and tropical
353 regions, like *S. trinodis*, must adapt to a broader range of environmental conditions
354 and pressures by increasing the physiological importance of phenolics to stress in
355 temperate regions and decreasing it in tropical ones.

356

357 The multiple regression models highlighted patterns of environmental factors likely to
358 be important influences on phenolic compound distribution in brown seaweeds. There
359 were significant positive relationships between seaweed phenolic concentrations and
360 salinity across study locations of Western Australia (Table 6a) as well as a subset of

361 individual species (Table 6b). High salinity can cause oxidative stress in seaweeds and
362 phenolic compounds act as antioxidative compounds to combat the elevated
363 production of reactive oxygen species (Pedersen 1984, Ragan and Glombitza 1986,
364 Zubia et al. 2007). Phenolic concentrations were also found in higher concentrations
365 at high salinities in *Ascophyllum nodosum* (Pedersen 1984). While salinity influences
366 phenolic concentrations, species salinity tolerances may also play a role. Phenolic
367 concentrations of some individual species were positively related to salinity in Shark
368 Bay (Table 6c). Seaweeds in tropical Shark Bay are tolerant to a wide a range of
369 salinity, which would explain why only *S. trinodis* and *S. decurrens* showed higher
370 phenolic concentrations at higher salinities (Figure 5) and not *Hormophysa*
371 *cuneiformis*. With a potentially broader salinity tolerance, a species like *H.*
372 *cuneiformis* would not need to increase cellular antioxidant activity through the
373 increased production of phenolic compounds.

374

375 We aimed to quantify differences phenolic concentrations of seaweeds throughout
376 coastal Western Australia. The coastline of Western Australia stretches almost 20,000
377 km (including offshore islands) and encompasses a variety of temperate and tropical
378 macroalgal assemblages (Kendrick et al. 1990; Huisman and Borowitzka 2003;
379 Wernberg et al. 2012; Huisman 2015). This resulted in sampling of different
380 macroalgal species at each study location with little species continuity (Table 2). A
381 few macroalgal species (*H. cuneiformis* and *S. trinodis*) were found at multiple
382 locations while all but two species were restricted to either temperate or tropical study
383 locations. Despite limited species continuity, the geographic scope of this project
384 highlighted the differences in distribution of phenolic compounds across temperate
385 and tropical ecosystems unlike other studies that encompassed only temperate

386 (Steinberg 1989; Tanniou et al 2014; Van Alstyne 1999)) or tropical (Fleury 1994;
387 Pavia and ÅBerg 1996; Steinberg 1986; Stiger et al 2004; Targett et al. 1992)
388 latitudes.

389

390 It is interesting that we found a similar distribution of phenolic compounds across
391 temperate and tropical latitudes in the distantly-related taxa of the Dictyotaceae and
392 Sargassaceae (Figure 4). Previously, no latitudinal patterns in phenolic concentration
393 were found in multiple seaweed taxa across 15 degrees of latitude from the northeast
394 Pacific region (Van Alstyne et al. 1999). Phenolic concentration variability was often
395 as high among geographically close sites as far ones. Additionally, variation in
396 phenolic concentration of *Ascophyllum nodosum* was small at large spatial scales but
397 large between at distances of a few meters (Pavia and ÅBerg 1996). Our findings
398 indicate strong environmental gradients may be present in nearshore Western
399 Australia that overcome the inherent variability in phenolic compounds between taxa
400 as well as geographically close study locations. Because of the range of temperate and
401 tropical conditions found among study locations (Table 1), it is difficult to determine
402 which of those conditions are most influential *in situ*, as responses of phenolic
403 compounds to environmental conditions are known to be species specific (Ragan and
404 Jensen 1978; Steinberg 1994; Tala et al. 2016). Nevertheless, regardless of which
405 factors ultimately drive phenolic concentrations, the differences in concentrations we
406 found are greater than the effects of study location variability.

407

408 We found no relationship between phenolic concentrations of seaweeds and depth in
409 individual species collected at temperate Jurien Bay across a depth gradient of 6.5 –
410 10.0 m or Port Gregory across 1.5 – 11.0 m depth (Table 6c). Many environmental

411 gradients are associated with increasing depth, including decreasing temperatures and
412 light levels (Nybakken 1993). Phenolic compounds are known to function as UV
413 protection and UVB radiation has been highlighted as a driver of phenolic
414 concentrations (Hay 1996; Pavia et al. 1997; Pavia and Brock 2000; Mannino et al.
415 2014). The sampling depths of this study (0 – 12 m) may experience similar UV-B
416 and temperature conditions in the clear warm waters of the Leeuwin Current
417 (Kendrick et al. 2009) without and impact to the phenolic concentrations of subtidal
418 seaweeds. However, we found very high phenolic concentrations in *Turbinaria*
419 *gracilis* and *H. cuneiformis* from tropical Cygnet Bay, which was the only intertidal
420 site sampled. Phenolic concentrations of *H. cuneiformis* were 3.5 times higher in
421 Cygnet Bay than the other three tropical locations sampled. Also, *T. gracilis* in
422 Cygnet Bay (Table 2) had two to four times more phenolic compounds than reported
423 elsewhere in the Indo-Pacific for *Turbinaria sp.* (Steinberg 1986; Van Alstyne and
424 Paul 1990; Targett et al. 1992; Stiger et al. 2004). Within Cygnet Bay, 10.0 m tidal
425 swings leave seaweeds in high intertidal lagoons exposed to high temperatures and
426 elevated UV-B radiation. Phenolic compounds absorb UV-B radiation within the
427 range of 195 to 265 nm commonly associated with photooxidation (Pavia et al. 1997;
428 Henry and Van Alstyne 2004). The ability of phenolic compounds to protect seaweeds
429 from photooxidation likely resulted in high concentrations of phenolics of tropical
430 intertidal seaweeds like those in Cygnet Bay compared to the tropical subtidal study
431 locations.

432

433 Herbivory is a known driver of both constitutive and inducible chemical defense traits
434 in marine seaweeds (Cronin and Hay 1996; Peckol et al. 1996; Pavia and Toth 2000;
435 Haavisto 2016). Herbivory pressures on seaweeds, as a result of more abundant large-

436 bodied herbivores in the tropics, are generally higher at lower latitudes in Western
437 Australia (Vanderklift and Kendrick 2004; Poore et al. 2012) and globally (Vermeij
438 1978; Gaines and Lubchenco 1982). While marine chemical defenses are generally
439 globally more abundant at lower latitudes (Hay and Fenical 1988), low phenolic
440 concentrations in tropical regions of Western Australia are paradoxical, considering
441 the suggested defensive role of phenolic compounds in tropical brown seaweeds
442 (Steinberg 1986; Steinberg 1989; Targett and Boettcher 1995). Temperate herbivores
443 are often deterred by high phenolic concentrations (Iken et al. 2009; McCarty and
444 Sotka 2013) but tropical herbivores readily consume phenol-rich algae (> 2 % dry
445 mass phenolics) (Steinberg and Paul 1990; Steinberg et al. 1991) indicating a lack of
446 deterrence by phenolic compounds. Herbivore deterrence from other compounds such
447 as non-polar metabolites (Steinberg and Paul 1990) or other bioactive compounds
448 (Amsler and Fairhead 2006) are possibly more important in this role for tropical
449 seaweeds. Phenolic compounds in tropical seaweeds of Western Australia may have
450 other primary functions such as light protection. Low palatability resulting in
451 herbivore deterrence is likely a secondary characteristic of these compounds.

452

453 While we do not know the specific phenolic species found in the taxa examined in this
454 study, the phenolic concentrations provided here should be a relative measure of the
455 reductive potential of the compounds, and thus of their bioactivity. Phenolic
456 compounds are diverse group of multifunctional molecules, which, among higher-
457 order taxa, differ in quantity, sizes, steric configurations, and numbers of hydroxyl
458 groups (Ragan and Glombitza 1986; Amsler and Fairhead 2006). The Folin-Ciocalteu
459 method measures extractable, non-bound phenolics by quantifying the reducing
460 ability of accessible hydroxyl groups in the compounds and thus can be used to

461 compare reactivities among broader taxonomic groups. Because the types and
462 structures of phenolic compounds within species are likely similar, the Folin-
463 Ciocalteu method can be used for intraspecific comparisons of phenolic
464 concentrations (Appel et al. 2001).

465

466 In conclusion, we found concentrations of phenolic compounds in brown seaweeds
467 were variable across multiple taxonomic and spatial scales. This study spanned both
468 temperate and tropical ecosystems and identified similar distribution patterns of
469 phenolic compounds among seaweed families across latitude. Our multiple regression
470 models related phenolic concentration to salinity and PAR, suggesting common
471 evolutionary mechanisms across taxa. Lower tropical phenolic concentrations found
472 here and in other studies (Steinberg 1986; Steinberg and Paul 1990) indicate a lack of
473 deterrence by phenolic compounds in tropical areas with high grazing pressures in
474 Western Australia. Additionally, the higher phenolic concentrations in Cygnet Bay
475 indicate phenolic compounds likely mediate elevated stresses associated with the
476 intertidal environment. Field experiments measuring the phenolic response of
477 seaweeds to environmental gradients will help identify the mechanistic drivers behind
478 the gradient correlations presented in this study. Expanding on models such as those
479 presented here will increase the understanding of the evolutionary pressures that shape
480 regional distributions of phenolic compounds.

481

482 **ACKNOWLEDGEMENTS**

483 This study was carried out as part of the PhD thesis research of the first author at the
484 University of Western Australia. We thank C. Tuckett, E. Gates, D. Bearham, R.
485 McCallum, K. van Hees for their help with field logistics and sample collection, L.

486 Mattio for help with species identification, M. Considine and G. Cawthray for
487 assistance in laboratory analyses and the anonymous reviewers that helped improve
488 this manuscript. Cygnet Bay sampling was funded by the Western Australian Marine
489 Science Institution (WAMSI) Kimberley Marine Research Program (project 2.2.4 to
490 G.A.K), and supported by the Bardi Jawi Ranger program. Jurien Bay and Port
491 Gregory sampling was supported by funding from the University of Western
492 Australia, The Hermon Slade Foundation and the Australian Research Council
493 awarded to T.W. An NHT-II Caring for our Country grant coordinated by WAMSI
494 awarded to G.A.K supported sampling in Shark Bay. A University of Western
495 Australia Postgraduate Student Research Grant awarded to D.H.vH funded laboratory
496 analyses.

497 **CONFLICT OF INTEREST:** Daniel H van Hees declares that he has no conflict of
498 interest. Ylva S Olsen declares that she has no conflict of interest. Thomas Wernberg
499 declares that he has no conflict of interest. Kathryn L Van Alstyne declares that she
500 has no conflict of interest. Gary A Kendrick declares that he has no conflict of
501 interest.

502 **ETHICAL APPROVAL:** This article does not contain any studies with animals
503 performed by any of the authors.

504 **AUTHOR CONTRIBUTIONS:** D.H.vH and G.A.K conceived of the ideas; D.H.vH
505 collected the data; D.H.vH and K.V.A ran the laboratory analyses; D.H.vH, K.L.V.A,
506 Y.O, T.W and G.A.K analysed the data; D.H.vH led the writing of the manuscript
507 with contributions from G.A.K, K.L.V.A., T.W., and Y.S.O.

508

509

510 **REFERENCES**

- 511 Amsler CD, Fairhead VA (2006) Defensive and sensory chemical ecology of
512 brown algae. *Adv Bot Res* 43:1–91.
- 513 Appel HM (1993) Phenolics in ecological interactions: The importance of
514 oxidation. *J Chem Ecol* 19:1521–1552. doi: 10.1007/BF00984895
- 515 Appel HM, Govenor HL, D’ascenzo M, Siska E, Schultz JC (2001) Limitations of
516 Folin assays of foliar phenolics in ecological studies. *J Chem Ecol* 27:761–
517 778. doi: 10.1023/A:1010306103643
- 518 Bennett S, Wernberg T, Connell SD, Hobday AJ, Johnson CR, Ploczanska ES
519 (2016) The “ Great Southern Reef ”: social, ecological and economic value
520 of Australia’s neglected kelp forests. *Mar Freshw Res* 67:47–56. doi:
521 <http://dx.doi.org/10.1071/MF15232>
- 522 Connan S, Goulard F, Stiger V, Deslandes E, Ar Gall E (2004) Interspecific and
523 temporal variation in phlorotannin levels in an assemblage of brown algae.
524 *Bot Mar* 47:410–416. doi: 10.1515/BOT.2004.057
- 525 Cronin G, Hay ME (1996) Induction of seaweed chemical defenses by amphipod
526 grazing. *Ecology* 77:2287–2301.
- 527 Estes JA, Steinberg PD (1988) Predation, herbivory and kelp evolution.
528 *Paleobiology* 14:19-36
- 529 Fleury BG, Kelecom A, Pereira RC, Teixeira VL (1994). Polyphenols, terpenes
530 and sterols in Brazilian Dictyotales and Fucales (Phaeophyta). *Botanica*
531 *Marina* 457–462
- 532 Gaines SD, Lubchenco J (1982) A unified approach to marine plant-herbivore
533 interactions. II. Biogeography. *Annu Rev Ecol Syst* 13:111–138.
- 534 Haavisto F (2016) Macroalgal Defenses Against Herbivory: Causes and

- 535 Consequences of Intraspecific Variation. University of Turku
- 536 Hay ME (1996) Marine chemical ecology: what's known and what's next? J Exp
537 Mar Bio Ecol 200:103–134. doi: 10.1016/S0022-0981(96)02659-7
- 538 Hay ME, Fenical W (1988) Marine plant-herbivore interactions: the ecology of
539 chemical defense. Annu Rev Ecol Syst 19:111–145.
- 540 Henry BE, Van Alstyne KL (2004) Effects of UV radiation on growth and
541 phlorotannins in *Fucus gardneri* (Phaeophyceae) juveniles and embryos. J
542 Phycol 40:527–533. doi: 10.1111/j.1529-8817.2004.03103.x
- 543 Huisman JM (2000) Marine Plants of Australia. University of Western Australia
544 Press, Nedlands
- 545 Huisman JM (2015) Algae of Australia: Marine Benthic Algae of North-western
546 Australia. CSIRO
- 547 Huisman JM, Borowitzka MA (2003) Marine benthic flora of the Dampier
548 Archipelago, Western Australia. Western Australia Museum, Perth
- 549 Iken K, Amsler CD, Amsler MO, McClintock JB, Baker BJ (2009) Field studies
550 on deterrent properties of phlorotannins in Antarctic brown algae. Bot Mar
551 52:547–557. doi: 10.1515/BOT.2009.071
- 552 Jormalainen V, Honkanen T (2004) Variation in natural selection for growth and
553 phlorotannins in the brown alga *Fucus vesiculosus*. J Evol Biol 17:807–820.
554 doi: 10.1111/j.1420-9101.2004.00715.x
- 555 Kendrick GA, Goldberg NA, Harvey ES, McDonald J (2009) Historical and
556 contemporary influence of the Leeuwin Current to the marine biota of the
557 Southern Western Australian Continental Shelf and the Recherche
558 Archipelago. J R Soc West Aust 92:211.
- 559 Kendrick GA, Huisman JM, Walker DI (1990) Benthic Macroalgae of Shark Bay,

- 560 Western Australia. Bot Mar 33:47–54. doi: 10.1515/botm.1990.33.1.47
- 561 Le Lann K, Ferret C, Vanmee E, Spangol C, Lhuillery M, Payri C, Stiger-
562 Pouvreau V (2012) Total phenolic, size-fractionated phenolics and
563 fucoxanthin content of tropical Sargassaceae (Fucales, Phaeophyceae) from
564 the South Pacific Ocean: Spatial and specific variability. Phycol Res 60:37–
565 50. doi: 10.1111/j.1440-1835.2011.00634.x
- 566 Lüder UH, Clayton MN (2004) Induction of phlorotannins in the brown
567 macroalga *Ecklonia radiata* (Laminariales, Phaeophyta) in response to
568 simulated herbivory - The first microscopic study. Planta 218:928–937. doi:
569 10.1007/s00425-003-1176-3
- 570 Mannino AM, Vaglica V, Oddo E (2014) Seasonal variation in total phenolic
571 content of *Dictyopteris polypodioides* (Dictyotaceae) and *Cystoseira*
572 *amentacea* (Sargassaceae) from the Sicilian coast. Flora Mediterr 24:39–50.
573 doi: 10.7320/FIMedit24.039
- 574 McCarty AT, Sotka EE (2013) Geographic variation in feeding preference of a
575 generalist herbivore: The importance of seaweed chemical defenses.
576 Oecologia 172:1071–1083. doi: 10.1007/s00442-012-2559-6
- 577 McGowran B, Li Q, Cann J, Padley D, McKirdy DM, Shafik S (1997)
578 Biogeographic impact of the Leewin Current in Southern Australia since the
579 late middle Eocene. Palaeogeogr Palaeoclimatol Palaeoecol 136:19–40.
- 580 Nybakken JW Marine Biology: An Ecological Approach. Harper Collins. New
581 York. 1993
- 582 Pavia H, ÅBerg P (1996) Spatial variation in polyphenolic content of
583 *Ascophyllum nodosum* (Fucales, Phaeophyta). Hydrobiologia 326–327:199–
584 203. doi: 10.1007/BF00047807

- 585 Pavia H, Brock E (2000) Extrinsic factors influencing phlorotannin production in
586 the brown alga *Ascophyllum nodosum*. Mar Ecol Prog Ser 193:285–294. doi:
587 10.3354/meps193285
- 588 Pavia H, Cervin G, Lindgren A, Åberg P (1997) Effects of UV-B radiation and
589 simulated herbivory on phlorotannins in the brown alga *Ascophyllum*
590 *nodosum*. Mar Ecol Prog Ser 157:139–146.
- 591 Pavia H, Toth GB (2000) Inducible Chemical Resistance to Herbivory in the
592 Brown Seaweed *Ascophyllum nodosum*. Ecology 81:3212–3225.
- 593 Peckol P, Krane JM, Yates JL (1996) Interactive effects of inducible defense and
594 resource availability on phlorotannins in the North Atlantic brown alga
595 *Fucus vesiculosus*. Mar Ecol Prog Ser 138:209–217.
- 596 Pedersen A (1984) Studies on phenol content and heavy metal uptake in fucoids.
597 Hydrobiologia 116–117:498–504. doi: 10.1007/BF00027732
- 598 Plouguerné E, Cesconetto C, Cruz CP, Pereira RC, da Gama BAP (2012) Within-
599 thallus variation in polyphenolic content and antifouling activity in
600 *Sargassum vulgare*. J Appl Phycol 24:1629–1635. doi: 10.1007/s10811-012-
601 9826-0
- 602 Poore AGB, Campbell AH, Coleman RA, Edgar GJ, Jormalainen V, Reynolds
603 PL, Sotka EE, Stachowicz JJ, Taylor RB, Vanderklift MA, Duffy JE (2012)
604 Global patterns in the impact of marine herbivores on benthic primary
605 producers. Ecol Lett 15:912–922. doi: 10.1111/j.1461-0248.2012.01804.x
- 606 R Core Team (2013) R. A language and environment for statistical computing. R
607 Foundation for Statistical Computing, Vienna, Austria. ISBN 3-900051-07-0,
608 URL <http://www.R-project.org/>
- 609 Ragan MA, Glombitza KW (1986) Phlorotannins, brown algal polyphenols. Prog

- 610 Phycol Res 4:129–241.
- 611 Ragan MA, Jensen A (1978) Quantitative studies on brown algal phenols. II.
612 Seasonal variation in polyphenol content of *Ascophyllum nodosum* (L.) Le
613 Jol. and *Fucus vesiculosus* (L.). J Exp Mar Bio Ecol 34:245–258. doi:
614 10.1016/S0022-0981(78)80006-9
- 615 Ridgway K (2009) CSIRO Oceans & Atmosphere - Hobart. CARS 2009-CSIRO
616 Atlas of Regional Seas.
- 617 Schoenwaelder MEA, Clayton MN (1999) The presence of phenolic compounds
618 in isolated cell walls of brown algae. Phycologia 38:161–166. doi:
619 10.2216/i0031-8884-38-3-161.1
- 620 Steinberg PD (1989) Biogeographical variation in brown algal polyphenolics and
621 other secondary metabolites: comparison between temperate Australasia and
622 North America. Oecologia 78:373–382.
- 623 Steinberg PD (1986) Chemical defenses and susceptibility of tropical marine
624 brown algae to herbivores. Oecologia 69:628–630.
- 625 Steinberg PD (1994) Lack of short-term induction of phlorotannins in the
626 Australasian brown algae *Ecklonia radiata* and *Sargassum vestitum*. Mar
627 Ecol Prog Ser 112:129–133.
- 628 Steinberg PD, Edyvane K, De Nys R, Birdsyr R, van Altena IA (1991) Lack of
629 avoidance of phenolic-rich brown algae by tropical herbivorous fishes. Mar
630 Biol 109:335–343.
- 631 Steinberg PD, Estes JA, Winter FC. (1995) Evolutionary consequences of food
632 chain length in kelp forest communities. Proc Nat Acad Sci 92(18):8145-
633 8148
- 634 Steinberg PD, Paul VJ (1990) Fish feeding and chemical defenses of tropical

- 635 brown algae in Western Australia. *Mar Ecol Prog Ser* 58:253–259. doi:
636 10.3354/meps058253
- 637 Stiger V, Deslandes E, Payri CE (2004) Phenolic contents of two brown algae,
638 *Turbinaria ornata* and *Sargassum mangarevense* on Tahiti (French
639 Polynesia): Interspecific, ontogenic and spatio-temporal variations. *Bot Mar*
640 47:402–409. doi: 10.1515/BOT.2004.058
- 641 Swanson AK, Druehl LD (2002) Induction, exudation and the UV protective role
642 of kelp phlorotannins. *Aquat Bot* 73:241–253.
- 643 Tala F, Velasquez M, Mansilla A, Macaya EC, Thiel M (2016) Latitudinal and
644 seasonal effects on short-term acclimation of floating kelp species from the
645 South-East Pacific. *J Exp Mar Bio Ecol* 483:31–41. doi:
646 10.1016/j.jembe.2016.06.003
- 647 Tanniou A, Vandanjon L, Incera M, Serrano Leon E, Husa V, Le Grand J, Nicolas
648 J-L, Poupart N, Kervarec N, Engelen A, Walsh R, Guerard F, Bourgougnon
649 N, Stiger-Pouvreau V (2014) Assessment of the spatial variability of
650 phenolic contents and associated bioactivities in the invasive alga *Sargassum*
651 *muticum* sampled along its European range from Norway to Portugal. *J Appl*
652 *Phycol* 26:1–16. doi: 10.1007/s10811-013-0198-x
- 653 Targett NM, Boettcher TE (1995) Tropical marine herbivore assimilation of
654 phenolic-rich plants. *Oecologia* 103:170–179.
- 655 Targett NM, Coen LD, Boettcher AA, Tanner CE (1992) Biogeographic
656 comparisons of marine algal polyphenolics: evidence against a latitudinal
657 trend. *Oecologia* 89:464–470. doi: 10.1007/BF00317150
- 658 Van Alstyne KL (1995) Comparison of three methods for quantifying brown algal
659 polyphenolic compounds. *J Chem Ecol* 21:45–58. doi: 10.1007/BF02033661

- 660 Van Alstyne KL, McCarthy JJ, Hustead CL, Duggins DO (1999) Geographic
661 variation in polyphenolic levels of northeastern Pacific kelps and rockweeds.
662 Mar Biol 133:371–379. doi: 10.1007/s002270050476
- 663 Van Alstyne KL, Paul VJ (1990) The Biogeography of Polyphenolic Compounds
664 in Marine Macroalgae: Temperate Brown Algal Defenses Deter Feeding by
665 Tropical Herbivorous Fishes. Oecologia 84:158–163. doi: 10.1007/S00442-
666 004-V
- 667 Vanderklift MA, Kendrick GA (2004) Variation in abundances of herbivorous
668 invertebrates in temperate subtidal rocky reef habitats. Mar Freshw Res
669 55:93–103. doi: 10.1071/MF03057
- 670 Vermeij GJ (1978) Biogeography and adaptation: patterns of marine life. Harvard
671 University Press, Cambridge
- 672 Walker DI (1985) Correlations between salinity and growth of the seagrass
673 *Amphibolis antarctica* (Labill.) Sonder & Aschers., in Shark Bay, Western
674 Australia, using a new method for measuring production rate. Aquat Bot
675 23:13–26.
- 676 Wernberg T, Russell BD, Moore PJ, Ling SD, Smale DA, Campbell A, Coleman
677 MA, Steinberg PD, Kendrick GA, Connell SD (2011) Impacts of climate
678 change in a global hotspot for temperate marine biodiversity and ocean
679 warming. J Exp Mar Bio Ecol 400:7–16. doi: 10.1016/j.jembe.2011.02.021
- 680 Wernberg T, Smale D, Tuya F, Thomsen MS, Langlois TJ, de Bettignies T,
681 Bennett S, Rousseaux CS (2012) An extreme climatic event alters marine
682 ecosystem structure in a global biodiversity hotspot. Nat Clim Chang 5:1–5.
683 doi: 10.1038/nclimate1627
- 684 Wernberg T, Thomsen MS, Connell SD, Russell BD, Waters JM, Zuccarello GC,

685 Kraft GT, Sanderson C, West JA, Gurgel CFD (2013) The footprint of
686 continental-scale ocean currents on the biogeography of seaweeds. PLoS
687 One 8:1–8. doi: 10.1371/journal.pone.0080168

688 Womersley HBS (1987) The Marine Benthic Flora of Southern Australia, Part II.
689 South Australian Government Printing Division, Adelaide

690 Zubia M, Robledo D, Freile-Pelegrin Y (2007) Antioxidant activities in tropical
691 marine macroalgae from the Yucatan Peninsula, Mexico. J Appl Phycol
692 19:449–458. doi: 10.1007/s10811-006-9152-5

693

694

695 Table 1. Physical characteristics of seven study locations in Western Australia and the macroalgal sampling
 696 summary of each location. Seaweeds at each location were collected at different numbers of sites. Species
 697 collected indicates the number of species collected at each location. Sampling sites crossed known
 698 environmental gradients.

Location	Longitude (E)	Latitude (S)	Region	Exposure	Substrate/ Habitat	Envir. Gradients Sampled	Depth (m)	No. of sites	No. species collected
Eagle Bay (EB)	115.1	-33.56	Temperate	North-facing, shallow	Sandy reef flat	None	2	1	6
Jurien Bay (JB)	115.09	-30.49	Temperate	West-facing, exposed	Complex coral/ limestone	Depth	6.5-10	3	8
Port Gregory (PG)	114.22	-28.18	Temperate	Lagoon and exposed reef	Sandstone boulders	Depth	1.5-11	3	6
Shark Bay (SB)	113.22	-25.62	Tropical	North-facing; shallow bay	Flat sand, seagrass	Salinity	1-2.5	7	5
Exmouth Gulf (EG)	114.2	-21.94	Tropical	North-facing	Flat sand/reef	None	1.5	1	6
Thevenard Island (TI)	115	-21.46	Tropical	Offshore Island; east side	Flat sand with coral heads	None	2	1	8
Cygnets Bay (CB)	122.88	-16.56	Tropical	East-facing; tidal	Intertidal reef flat	None	Intertidal	3	7

699
700

701

702 Table 2. Phenolic concentrations (% DM, means +/- SE) of brown seaweeds
 703 collected at seven study locations along the coast of Western Australia. N is the
 704 number of tissue samples analysed. Study locations are abbreviated as follows:
 705 Cygnet Bay (CB), Jurien Bay (JB), Port Gregory (PG), Shark Bay (SB), Exmouth
 706 Gulf (EG), Thevenard Island (TI), Cygnet Bay (CB).
 707

Order	Family	Species	Location	N	Total Phenolics (%DM)		
Fucales	Dictyotaceae	<i>Dictyopteris muelleri</i>	JB	5	0.48 +/- 0.03		
		<i>Dictyota australis</i>	TI	5	1.63 +/- 0.34		
		<i>Dictyota ciliolata</i>	PG	3	0.95 +/- 0.45		
		<i>Dictyota cylanica</i>	TI	4	0.35 +/- 0.03		
			EG	5	0.35 +/- 0.03		
		<i>Dictyota naevosa</i>	PG	3	1.06 +/- 0.43		
		<i>Dictyota sp</i>	PG	2	0.43 +/- 0.03		
		<i>Dilophus sp</i>	JB	3	1.11 +/- 0.14		
		<i>Distromium sp</i>	PG	12	2.24 +/- 0.38		
			JB	4	4.40 +/- 0.72		
		<i>Lobophora variegata</i>	EG	10	0.31 +/- 0.01		
			EB	5	8.73 +/- 1.05		
		<i>Padina sp.</i>	CB	6	2.18 +/- 0.17		
			TI	4	1.73 +/- 0.03		
			EG	4	1.95 +/- 0.11		
		<i>Spatoglossum macrodontum</i>	JB	3	1.81 +/- 0.35		
		<i>Zonaria turneriana</i>	JB	5	4.78 +/- 0.39		
		Sargassaceae		<i>Cystophora #1</i>	EB	4	9.36 +/- 0.68
				<i>Cystophora grevillei</i>	EB	3	10.99 +/- 2.63
				<i>Hormophysa cuneiformis</i>	CB	12	2.59 +/- 0.19
					EG	5	0.72 +/- 0.10
SB	24				0.76 +/- 0.12		
<i>Myriodesma serrulata</i>	JB			5	1.17 +/- 0.15		
<i>Sargassopsis decurrens</i>	TI			6	0.51 +/- 0.07		
	SB			20	0.99 +/- 0.12		
<i>Sargassum ligulatum</i>	CB			8	1.13 +/- 0.14		
	TI			5	0.25 +/- 0.03		
	EG			6	0.33 +/- 0.04		
<i>Sargassum linearifolium</i>	EB			6	4.16 +/- 0.75		
<i>Sargassum marginatum</i>	TI			6	0.70 +/- 0.11		
<i>Sargassum olygocystin</i>	PG			5	0.28 +/- 0.11		
<i>Sargassum paradoxum</i>	PG			19	1.28 +/- 0.10		
	JB			5	2.94 +/- 0.27		
<i>Sargassum polycistum</i>	CB			8	1.42 +/- 0.14		
<i>Sargassum polyphyllum</i>	EG			5	0.63 +/- 0.08		
<i>Sargassum rasta</i>	CB			6	0.81 +/- 0.05		
<i>Sargassum unknown #4</i>	TI			5	0.31 +/- 0.07		
<i>Scaberia agardhii</i>	EB			5	4.73 +/- 0.65		
	CB			6	1.14 +/- 0.15		
<i>Sirophysalis trinodis</i>	TI			3	0.87 +/- 0.19		
	EG			5	0.55 +/- 0.12		
	SB	13	1.05 +/- 0.34				
	EB	3	3.56 +/- 0.92				
<i>Turbinaria gracilis</i>	CB	15	2.96 +/- 0.28				
Laminariales	Lessonaceae						
		<i>Ecklonia radiata</i>	JB	13	3.37 +/- 0.28		
Sporochnales	Sporochnaceae						
		<i>Sporochnus moorei</i>	SB	4	1.15 +/- 0.32		

711 Table 3. Values of environmental variables extracted from MODIS satellite data for the seven nearshore study
 712 locations of Western Australia. Values represent seasonal composites for the austral fall of 2015. Chlorophyll a
 713 values are reported in mg m^{-3} , diffuse attenuation at 490 nm m^{-1} , photosynthetically active radiation in Einstein m^{-2}
 714 day^{-1} , salinity, sea surface temperature in $^{\circ}\text{C}$.

Factor	Eagle Bay	Jurien Bay	Port Gregory	Shark Bay	Exmouth Gulf	Thevenard Island	Cygnets Bay
Chlorophyll a (mg m^{-3})	0.43	1.69	1.15	3.29	3.06	4.25	1.11
Diffuse attenuation (490nm m^{-1})	0.06	0.15	0.11	0.24	0.22	0.29	0.11
Photosynthetically active radiation (Einstein $\text{m}^{-2} \text{day}^{-1}$)	26.75	29.65	30.98	32.01	35.78	34.41	43.48
Salinity	35.76	35.66	35.66	39 - 53.5	35.05	35.15	35.07
Sea surface temperature ($^{\circ}\text{C}$)	26.30	22.70	22.00	25.40	26.30	26.60	28.51

715
716

717

718

719 Table 4. Results of ANOVAs comparing phenolic concentrations of study
 720 locations and sample sites to the mean phenolic concentration of a) the study
 721 location and of b) the mean concentrations of each family. Tests were performed
 722 on fourth-root transformed data. All seven study locations were used in each
 723 statistical test.

Taxa	Factor	DF	MS	F	P
a) Assemblage	Location	6	0.29	16.77	< 0.001
	Location(Site)	16	0.28	0.62	0.85
	Error	45	0.13		
b) Family	Family	3	0.02	2.86	0.04
	Location	6	0.27	16.63	<0.001
	Location(Site)	16	0.03	0.68	0.79
	Error	42	0.12		

724
725

726

727

728

729

730 Table 5. ANOVA results comparing the phenolic concentrations of a) brown
 731 seaweed species across multiple study locations and b) species between multiple
 732 sites within a given study location in Western Australia. Tests were performed on
 733 fourth-root transformed data. The study locations used in each variance test are
 734 noted under “Location”.

Species	Factor	DF	MS	F	P	Location
a)						
<i>Hormophysa cuneiformis</i>	Location	2	0.56	33.7	< 0.001	SB, EG, CB
	Error	38	0.02			
<i>Sargassum ligulatum</i>	Location	2	0.19	30.75	< 0.001	EG, TI, CB
	Error	16	0.01			
<i>Sirophysalis trinodis</i>	Location	4	0.14	4.44	0.008	EB, SB, EG, TI, CB
	Error	25	0.03			
b)						
<i>Ecklonia radiata</i>	Site	2	0.001	0.01	0.98	JB
	Error	10	0.01			
<i>Sargassum paradoxum</i>	Site	3	0.09	9.99	<0.001	PG
	Error	20	0.01			
<i>Hormophysa cuneiformis</i>	Site	4	0.06	4.1	0.01	SB
	Error	19	0.02			
<i>Sargassopsis decurrens</i>	Site	2	0.03	2.65	0.11	SB
	Error	12	0.01			
<i>Sirophysalis trinoids</i>	Site	2	0.22	12.96	0.002	SB
	Error	10	0.02			
<i>Hormophysa cuneiformis</i>	Site	1	0.03	6.99	0.02	CB
	Error	10	0.005			
<i>Turbinaria gracilis</i>	Site	2	0.049	4.97	0.03	CB
	Error	12	0.001			

735
736

737

738

739

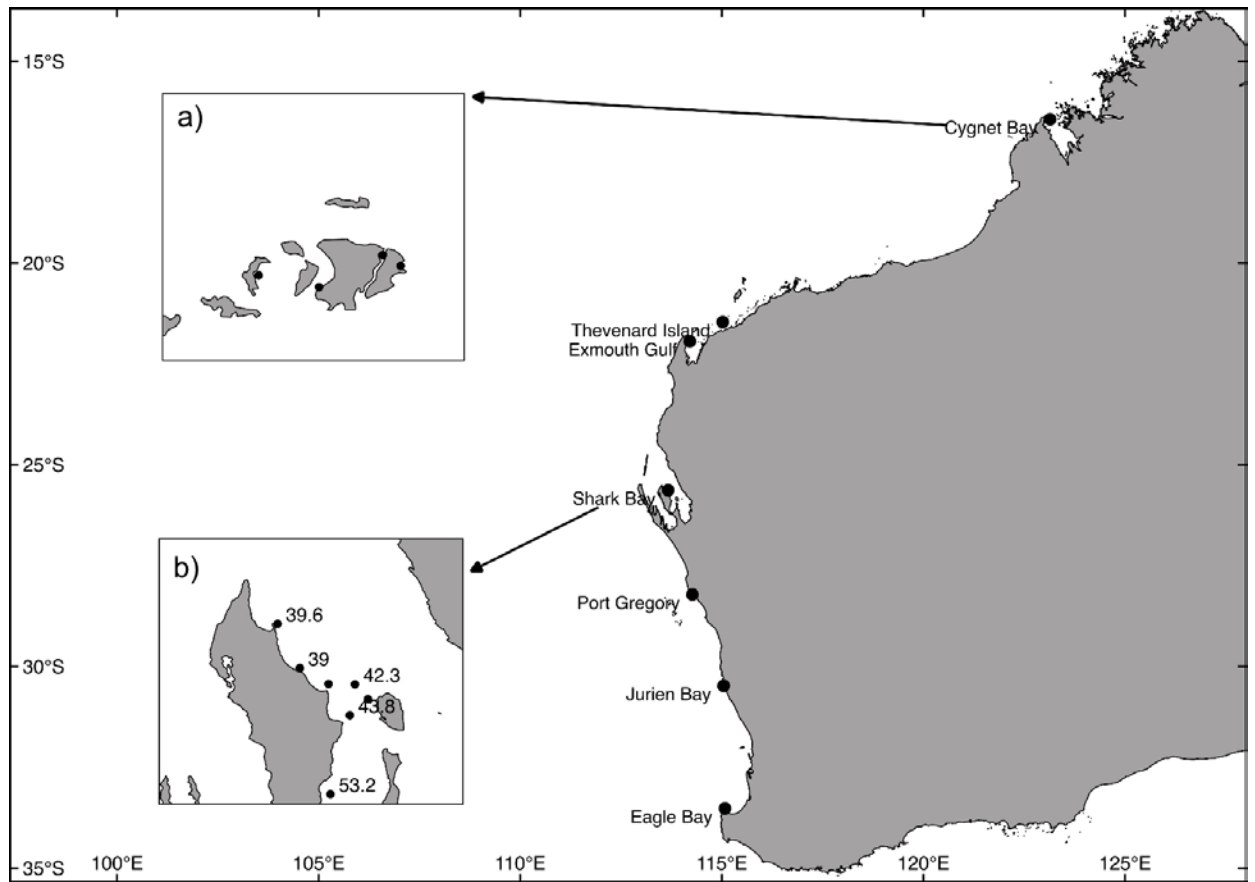
740

741

742

743 Table 6. Results from multiple regression analyses comparing the relationships of
 744 abiotic conditions to brown seaweed phenolic concentrations of (a) study
 745 locations and (b) individual families and species. Regressions comparing phenolic
 746 concentrations of individual species across local environmental gradients within
 747 sites are shown in (c). The multiple regression model compared phenolic
 748 concentration to: diffuse attenuation, PAR, salinity and sea surface temperature.
 749 Each regression used either all or a subset of the study locations as indicated.

Taxa	Regression	Factor	p	R ²	Partial η^2	Relationship	Location(s)
a)							
All Species	Single	Latitude	< 0.001	0.49			EB, JB, PG, SB, EG, TI
All Species	Multiple		< 0.001	0.57			EB, JB, PG, SB, EG, TI
		Attenuation			0.003	Negative	
		PAR			0.18	Negative	
		Salinity			0.15	Positive	
		SST			0.0001	Negative	
b)							
Dictyotaceae	Multiple		< 0.001	0.49			EB, JB, PG, SB, EG, TI
		Attenuation			0.036	Negative	
		PAR			0.109	Negative	
		Salinity			0.14	Positive	
		SST			0.05	Negative	
Sargassaceae	Multiple		< 0.001	0.65			EB, JB, PG, SB, EG, TI
		Attenuation			0.03	Negative	
		PAR			0.27	Negative	
		Salinity			0.21	Positive	
		SST			0.009	Negative	
<i>Hormophysa cuneiformis</i>	Multiple		< 0.001	0.64			CB, EG, SB
		Attenuation			-1.33 e ⁻¹⁴	Negative	
		PAR			0	-	
		Salinity			0	-	
		SST			-5.26 e ⁻¹⁶	Positive	
<i>Sargassum ligulatum</i>	Multiple		< 0.001	0.79			CB, EG, TI
		Attenuation			-4.6 e ⁻¹⁵	Negative	
		PAR			0	-	
		Salinity			0	-	
		SST			-4.18 e ⁻¹⁵	Positive	
<i>Sirophysis trinodis</i>	Multiple		0.007	0.42			CB, EB, EG, SB, TI
		Attenuation			0.067	Negative	
		PAR			0.093	Negative	
		Salinity			0.095	Positive	
		SST			0.1	Negative	
c)							
<i>Hormophysa cuneiformis</i>	Single	Salinity	0.18	0.08		Pos	SB
<i>Sargassopsis decurrens</i>	Single	Salinity	0.03	0.3		Pos	SB
<i>Sirophysis trinodis</i>	Single	Salinity	0.02	0.29		Pos	SB
<i>Ecklonia radiata</i>	Single	Depth	0.88	0.02		-	JB
<i>Sargassum paradoxum</i>	Single	Depth	0.08	0.13		Neg	JB
<i>Sargassum paradoxum</i>	Single	Depth	0.23	0.16		Neg	PG

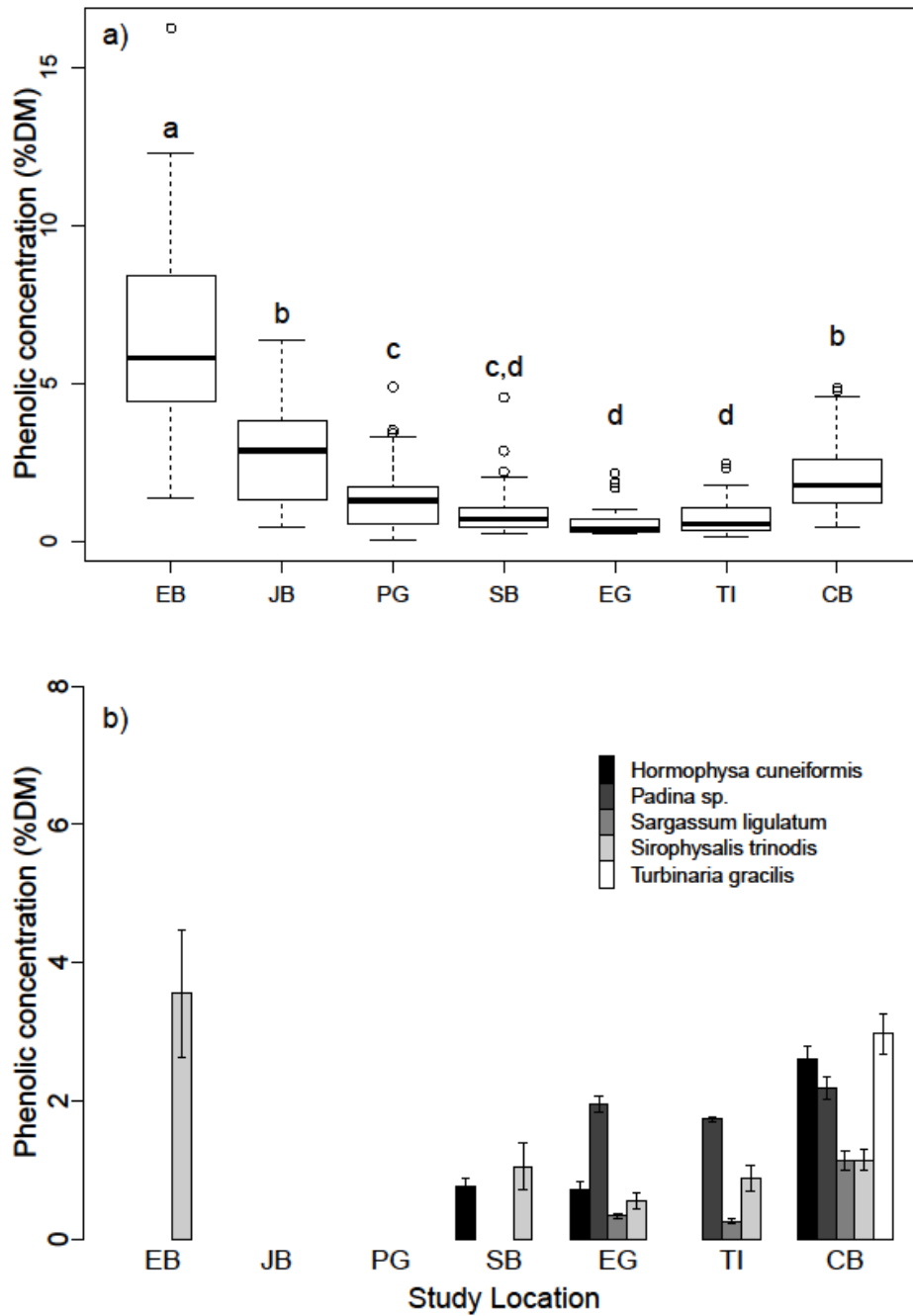


752

753 Figure 1. Map of study locations along coastal Western Australia. The inset map
 754 of Cygnet Bay (a) shows the locations of sample sites around Sunday Island. Inset
 755 map of Shark Bay (b) shows sampling sites along with the associated salinity of
 756 each site.

757

758



759

760

761

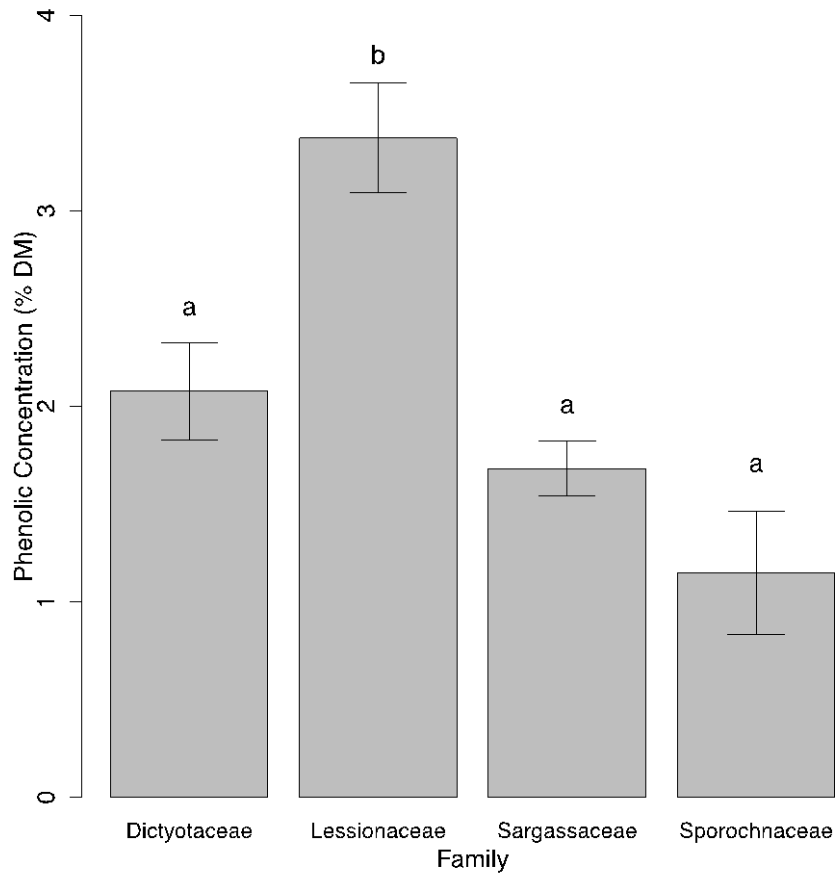
762

763

764

765

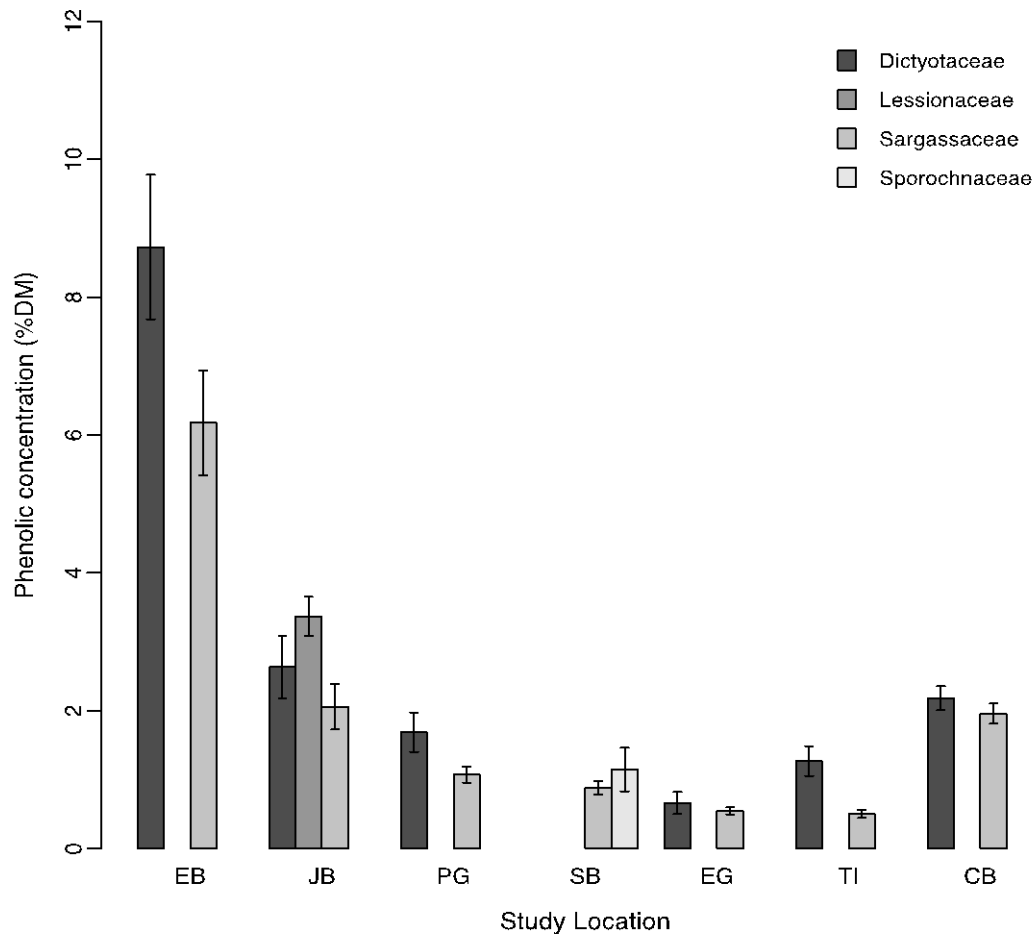
Figure 2. Mean phenolic concentration (as % dry mass) of a) the dominant brown seaweeds (as a composite of all species collected) and b) individual species found across study locations along nearshore Western Australia. Values shown are means \pm SE. Collection locations are ordered by high to low latitude from left to right. Letters indicate significant differences between study locations (from Tukey's HSD).



766

767 Figure 3. Phenolic concentrations (as % dry mass) of brown seaweed families
768 collected at seven study locations across nearshore Western Australia. Each bar
769 represents the mean phenolic concentration averaged over the seven locations.
770 Values are means \pm SE. Letters indicate significant differences between families
771 (from Tukey's HSD).

772



773

774 Figure 4. Phenolic concentrations (as % dry mass) of macroalgal families

775 collected at seven nearshore study locations along Western Australia. Study

776 locations are organized highest to lowest latitude, left to right. Values shown are

777 means \pm SE.

778

779

780

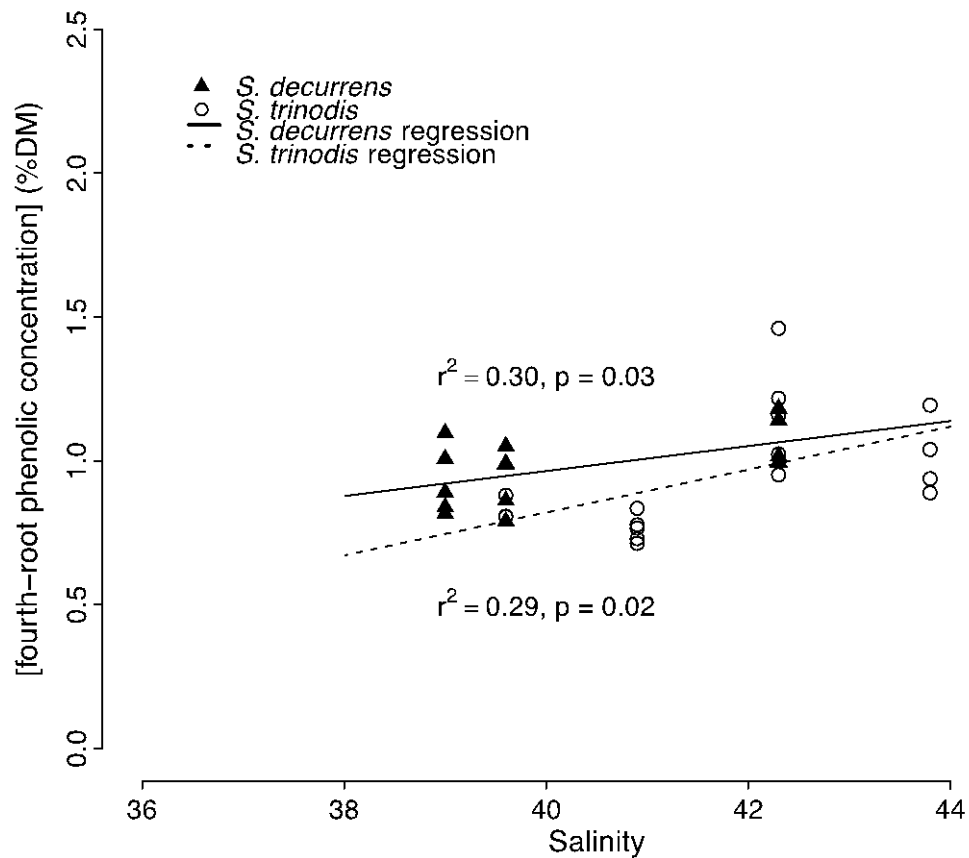
781

782

783

784

785



786

787

788

789

790

791

792

Figure 5. Linear regressions comparing fourth-root transformed phenolic concentrations of *Sargassopsis decurrens* (triangles) and *Sirophysis trinodis* (circles) to salinity in Shark Bay, Western Australia. Regression lines indicate the relationship between salinity and phenolics for *S. decurrens* (solid) and *S. trinodis* (dashed).