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2 Research Paper

3

4 **TITLE**

5 Interpreting vegetation change in tropical arid ecosystems from sediment molecular fossils and
6 their stable isotope compositions: A baseline study from the Pilbara region of northwest Australia

7

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24 **HIGHLIGHTS (max 85 characters)**

- 25 • Biomarkers and *n*-alkane $\delta^{13}\text{C}$ were extracted from sediment with total C = 0.4–1.4%
- 26 • Distribution of biomarkers reflected a decrease of aridity in the last ~2000 years
- 27 • Arid based Bayesian model of *n*-alkanes $\delta^{13}\text{C}$ suggested increase of riparian C_3 input
- 28 • Provided interpretive framework for reconstructions in tropical arid ecosystems

29

30 **KEY WORDS**

31 Biomarkers; CSIA $\delta^{13}\text{C}$; *n*-alkanes; organic matter; Pilbara; *Triodia*

32 **ABSTRACT**

33 Detection of source diagnostic molecular fossils (biomarkers) within sediments can provide
34 valuable insights into the vegetation and climates of past environments. However, hot and arid
35 regions offer particularly challenging interpretive frameworks for reconstructing past conditions
36 because baseline data are scarce, sedimentary organic matter contents are generally very low and
37 in the inland tropics in particular, sediments are also often subject to periods of flooding and
38 drought. However, there have been relatively few assessments of whether it is possible for
39 biomarkers to be extracted in sufficient quantities from arid sediments to allow interpretation of
40 past changes in vegetation and climate. Here, we measured the amounts, molecular distribution
41 and $\delta^{13}\text{C}$ values of organic compounds extracted from a multi-dated (^{210}Pb , ^{137}Cs and ^{14}C) late
42 Holocene sediment record from the Fortescue Marsh (Pilbara, northwest Australia) to investigate
43 catchment vegetation history in relation to hydroclimatic change over recent millennia. The low
44 total carbon (TC) content (<1.4 %) was a major challenge for the molecular analyses over the
45 ~2000 years old sequence. Nevertheless, they revealed that the dominant hydrocarbon features
46 (e.g., long chain *n*-alkanes) indicative of terrestrial plants (e.g., C_4 grasses; riparian and other C_3
47 plants) encompassed the last ~1300 yrs and that low abundance of products from aquatic sources
48 (e.g., *n*- C_{17}) were detected in the uppermost sediments only when permanently inundated
49 conditions prevailed (recent decades). Similarly, the lower $\delta^{13}\text{C}$ values (i.e., a difference of -2.3 ‰)
50 of long chain *n*-alkanes in upper sediments reflected a vegetation response to the emergence of
51 wetter conditions through the late Holocene in the region. Based on the diverging dominant
52 source contributions obtained from the molecular distributions and arid based Bayesian mixing
53 model ($\delta^{13}\text{C}$ of *n*- C_{27-33} alkanes) results, less arid conditions may have favoured the input of ^{13}C
54 depleted *n*-alkanes from the *Eucalyptus* (C_3) dominant riparian vegetation. The deepest sediments
55 (<CE 700) however, had a TC content of <0.4 %, and no organic compounds were detected,
56 consistent with local and regional records of hyperarid conditions. These results demonstrate that

57 *n*-alkanes can provide a molecular and stable isotopic fingerprint of important - and perhaps
58 underappreciated - ecological processes in modern tropical arid environments for future
59 paleoclimate investigations.

60 **ABBREVIATIONS**

61

62 CE Common Era

63 CPI Carbon preference index

64 CSIA Compound-specific stable isotope analysis

65 GC-MS Gas chromatography-Mass spectrometry

66 IRMS Isotope ratio mass spectrometry

67 OM Organic matter

68 TC Total carbon

69

70

71 **1. INTRODUCTION**

72 Lake sediment records can extend temporal scales of hydrologic reconstruction to millennial and
73 longer time frames. For example, periods of drought can be reconstructed from changes in lake
74 level and salinity using microfossil assemblages (e.g., diatoms) and geochemical or physical
75 parameters of the sediment (e.g., Wolff et al. 2011; Barr et al. 2014). These paleolimnological
76 methods have most commonly been applied to closed-basin lakes in arid or semiarid landscapes,
77 in which changes in lake level and salinity are closely related to shifts in hydrologic balance (e.g.,
78 Stager et al. 2013). However, deep or permanent lakes are often lacking in many arid
79 environments, resulting in limited applicability of many proxies due to preservation issues, notably
80 for reconstructing inputs of organic matter to lakes that can reflect hydroclimate conditions.
81 Organic geochemical studies of depositional environments have provided molecular evidence
82 (biomarkers) of changing hydrology under various climates, including from the studies of lakes
83 (Leng and Henderson 2013; Sun et al. 2013; Atahan et al., 2015), marine settings (e.g., Castañeda
84 et al. 2009b; Dubois et al. 2014; Pagès et al. 2014), and coastal salt-marshes and lagoons (e.g.,

85 Volkman et al. 2007; McKirdy et al. 2010; Tulipani et al., 2014). The sedimentary distribution of
86 biomarkers and their stable carbon and hydrogen isotope compositions (i.e., $\delta^{13}\text{C}$ and $\delta^2\text{H}$ at
87 compound specific level) have been widely used to investigate physiological or ecological
88 community responses to shifts in lake salinity and water source (Romero-Viana et al. 2012; Sachse
89 et al. 2012; van Soelen et al. 2013). Variation in the $\delta^{13}\text{C}$ values of specific compounds, e.g., long
90 chain *n*-alkanes representative of terrestrial vegetation, have also indirectly helped resolve past
91 hydrological changes via their sensitivity to moisture availability in the catchment, drought stress
92 or vegetation shifts (Castañeda et al. 2007; 2009a; Kristen et al. 2010; Tipple and Pagani 2010; Sun
93 et al. 2013). Consequently, using biomarkers and their stable isotopes ratios to establish
94 paleoenvironmental records for still underrepresented regions is of great interest.

95

96 Tropical arid ecosystems have proved especially challenging to paleolimnology and organic
97 geochemical characterisation of sediments. These regions experience high seasonal and inter-
98 annual hydroclimatic variability, typified by periods of prolonged droughts interspersed with
99 occasional intense flooding (e.g., Rouillard et al. 2015; van Etten 2009). First, a general lack of
100 water, which is often coupled with low nutrient availability, severely limits biological production
101 and in turn the deposition of organic matter (OM) to sediments (e.g, Huxman et al. 2004; Snyder
102 and Tartowski 2006; Collins et al. 2008; McIntyre et al. 2009a, b). In Australian soils of tropical (i.e.,
103 hot) arid environments in particular, the preservation of OM is further compromised by major
104 biological reworking by termites and other fauna (e.g., Holt et al. 1987; Chen et al. 2003; Austin et
105 al. 2004) and the high frequency of fires which can accelerate the loss of organic volatiles that
106 might otherwise form terrestrial detritus (Ford et al. 2007). Intense episodic flooding can lead to
107 increased mobilisation and dislocation of OM from the catchment to lakes followed by pulses of
108 high organic production in the catchment and aquatic network, resulting in non-linear OM
109 availability and deposition patterns (Battin et al. 2008; Reid et al. 2011; Puttock et al. 2012).

110 Shallow lakes mean that prolonged drought further exposes dried sediment to changing redox
111 conditions and deflation, which can also cause hiatus of organic sedimentary records in many
112 regions of the world (Verschuren et al. 1999; Bessems et al. 2008; Argus et al. 2014, 2015). Given
113 these challenges, there have been few attempts to characterise such sediments in northwest
114 Australia, a region where very limited knowledge of the Holocene environmental history is
115 available to understand long-term ecosystem processes and resilience to ongoing climate change
116 (Gergis et al. 2014).

117

118 Plant molecular $\delta^2\text{H}$ signals are normally considered indicative of the isotope ratio of source water
119 and associated hydrology (Tippie and Pagani 2010; Romero-Viana et al. 2012; Sachse et al. 2012;
120 van Soelen et al. 2013). However, the large evaporative shifts in lake volumes in arid and semi-arid
121 regions generally result in salinity fluctuations that in turn affect the $\delta^2\text{H}$ fractionation between
122 organisms and their source water (Sachse and Sachs 2008). In addition, plants make physiological
123 adjustments to both changing water availability and salinity, which can also result in further
124 fractionation (Zhou et al. 2011; Sachse et al. 2012), making plant molecular $\delta^2\text{H}$ less reliable than
125 in more temperate or non-saline ecosystems (e.g., Duan et al., 2014). Consequently, plant
126 molecular $\delta^{13}\text{C}$ compositions of sediments, particularly of *n*-alkanes, may be more suitable for hot
127 arid environments.

128

129 The composition and distribution of plant communities in arid ecosystems is largely driven by the
130 availability of water as well as other flooding or waterlogging-related conditions, such as soil
131 redox, salinity, substrate and nutrient availability and the accumulation of toxic ions (Schwinning
132 et al. 2004; Reid et al. 2011; Argus et al. 2014). Different hydrological regimes will favour plants of
133 specific functional types (e.g., C_3 versus C_4 photosynthetic pathways, which are readily
134 distinguished by their $\delta^{13}\text{C}$ signatures) and tolerance to drought, waterlogging or salinity (Huxman

135 et al. 2004; Snyder and Tartowski 2006). Prolonged aridity can thus progressively impact
136 vegetation assemblages at the generation level over long timescales, generally favouring C₄
137 grasses across catchments and chenopods (often a mix of CAM, C₄ and C₃ species) in saline
138 wetland areas (Sage et al. 2011). The ¹³C enrichment of C₃ plant tissue can also be driven by water
139 stress within a matter of minutes (Cernusak et al. 2013). The dynamics of plant water stress over
140 periods of decades to centuries may also be recorded in the wood chemistry of trees (e.g., Cullen
141 et al. 2008; Qin et al. 2015) or at the ecosystem level in soils and lake sediment (Krull et al. 2005,
142 2006; Diefendorf et al. 2011; Codron et al. 2013). Plants in arid environments also produce
143 relatively high amounts of cuticular waxes as a protection against high ultraviolet light and water
144 loss; cuticular waxes may thus represent more promising molecular targets of organic lean dryland
145 sediments (Hoffmann et al. 2013).

146

147 Cuticular-derived *n*-alkanes (Eglinton and Hamilton 1967) are typically some of the more abundant
148 products from the organic sediments of lakes (Schmidt et al. 2011), so are attractive analytes for
149 compound specific isotopic analysis (CSIA). Characterisation of the molecular and stable isotopic
150 abundances of *n*-alkanes has become an important paleoecological strategy to assess changes in
151 hydrological status at catchment and continent scales (Castañeda and Schouten 2011; Sachse et
152 al. 2012; Leng and Henderson 2013; Dubois et al. 2014). In recent years, the $\delta^{13}\text{C}$ values of *n*-
153 alkanes in sediments have unlocked a wealth of paleoenvironmental information from regions and
154 sites that had previously been difficult, including assessment of historical change in catchment
155 land-use, eutrophication studies and shifts in vegetation composition in response to climate (e.g.,
156 Castañeda and Schouten 2011; Fang et al. 2014). However, as with other relatively new molecular
157 tools in paleolimnology, our capacity to make inferences of paleo-flora from biomarkers is limited
158 by uncertainties in source, transportation and preservation of OM from senescing plants (Birks
159 and Birks 2016). A first step is to improve the baseline of available biomarker datasets and their

160 stable isotopic compositions from arid zone sediments with independently described
161 paleoenvironmental histories.
162
163 Broadly, our study investigated the potential of organic molecular fossils and $\delta^{13}\text{C}$ values of plant
164 waxes of lake sediments for reconstructing catchment vegetation history and hydroclimatic
165 change in the arid subtropical Pilbara region of inland northwest Australia. Previously, we
166 established centennial and millennial flood records for the Pilbara based on historical records as
167 well as physical and geochemical analyses of sediments at the Fortescue Marsh (Rouillard et al.
168 2015; Rouillard et al. 2016). Here, we used a replicate core of this sediment sequence to
169 determine if: i) organic geochemical information can be extracted to identify the major sources of
170 sedimentary OM; ii) $\delta^{13}\text{C}$ values of *n*-alkanes reflect the major vegetation types in the catchment
171 (C_3 and C_4); and iii) the molecular fossils and their $\delta^{13}\text{C}$ values from sedimentary OM can be related
172 to change in hydrology through the late Holocene in the region. Our analysis is considered within
173 an interpretive framework that accounts for the specificities and challenges of extracting
174 palaeoclimate information from organic geochemical indicators in sediments developed in tropical
175 arid ecosystems.

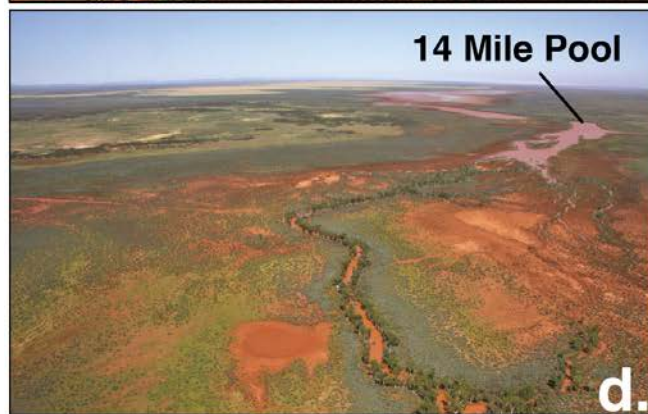
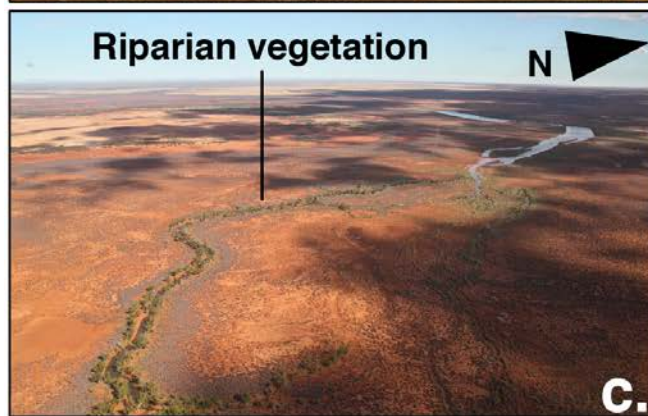
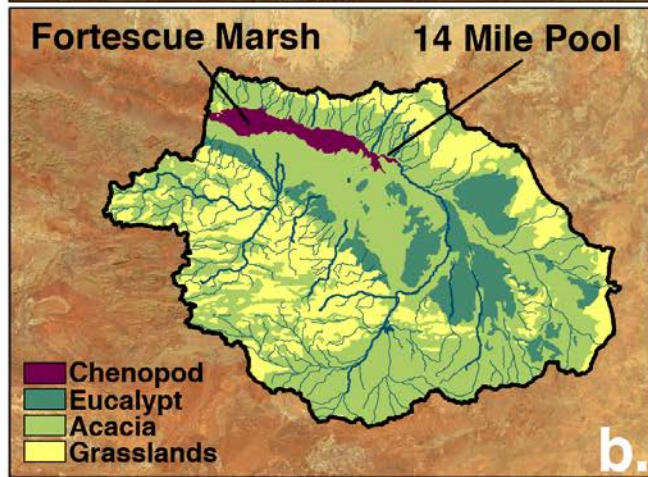
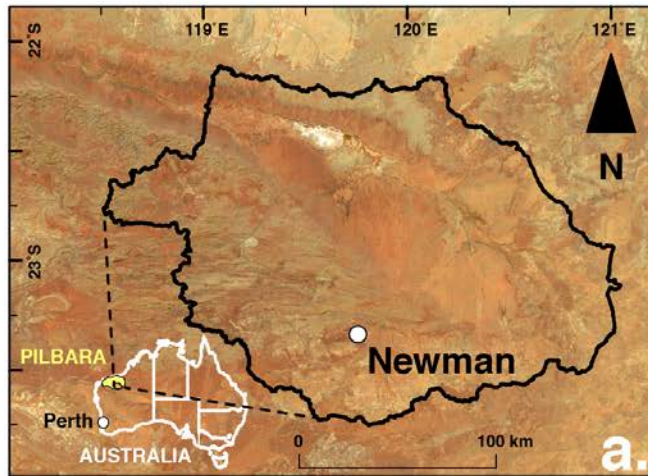
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177 **2. METHODS**

178 **2.1 Study site**

179 The study region and sampling is described in detail in Rouillard et al. (2016 in review). Briefly,
180 replicated ~60-cm sediment cores (FOR106C2 and FOR106C3) were obtained from the deepest
181 section of the 14 Mile Pool (22.6°S, 119.9°E) of the Fortescue Marsh in the arid Pilbara region of
182 northwest Australia (Fig. 1). The pool is the most perennial water feature of the Fortescue Marsh,
183 which acts as a terminal lake for the Upper Fortescue River catchment (31,000 km²) (Skrzypek et.
184 al. 2013). The Pilbara experiences highly episodic large hydrological events with high inter-annual

185 variability, i.e., creeks and rivers are generally dry with seasonal floods commonly occurring during
186 the austral summer contributing brief periods of high flow (www.water.wa.gov.au). Evaporation is
187 high in the eastern Pilbara ($\sim 2800\text{--}3200\text{ mm}\cdot\text{yr}^{-1}$) and greatly exceeds average rainfall (290
188 $\text{mm}\cdot\text{yr}^{-1}$; www.bom.gov.au/climate/data). Flooding, drought and fire are the dominant and
189 naturally occurring ecological disturbances in the region (McKenzie et al. 2009); variability in the
190 seasonal and inter-annual hydrological cycle is extreme (Kennard et al. 2010; Rouillard et al. 2015)
191 and likely influences the fire regime through feedback interactions (e.g., O'Donnell et al. 2011).



192

193

194 **Figure 1: a–b)** Upper Fortescue River catchment delineated (solid black line) on a True-colour
195 Aqua MODIS image (5, Jan 2005; NASA 2005), including b) a vegetation map with Chenopod
196 (Shrublands, Sampire Shrubland and Forblands), Eucalypt (Woodlands; Open Woodlands), Acacia
197 (Shrublands; Open Woodlands; Forests and Woodlands) and Grasslands (Hummock (*Triodia* spp);
198 Tussock); and **c–d)** aerial photographs (flowing East–West, the Upper Fortescue River, the 14 Mile
199 Pool and the Fortescue Marsh) of the **c)** riparian vegetation remaining (green) during dry
200 conditions (Jul, 2010; Photograph: Clive Taylor) and **d)** the greening effect of large rains during the
201 wet season (Feb 2012, one month after Severe Tropical Cyclone Heidi).

202 (1-COLUMN FITTING IMAGE)

203

204 Vegetation across the hills and broad plains of the wider catchment is mainly shrub-steppe,
205 dominated by C₄ grasses, especially spinifex (*Triodia* spp.) grasses and *Acacia* trees and shrubs (C₃).
206 More permanent watercourses, including 14 Mile Pool, maintain a riparian fringe dominated by
207 the trees *Eucalyptus camaldulensis* and *E. victrix* (C₃) (Fig. 1b, d; Beard 1975). Buffel grass
208 (*Cenchrus ciliaris*—C₄) is an invasive perennial grass introduced in the 1920s that was also aerial
209 seeded across the catchment in the 1960s to provide higher quality feed for cattle and sheep
210 grazing; this grass is abundant in the floodplains of sections of the Upper Fortescue River
211 (McKenzie et al. 2009). The Marsh downstream from 14 Mile Pool is dominated by chenopod
212 shrubs, primarily *Tecticornia* species (Beard 1975) that have been attributed both C₃ and C₄
213 pathways (Carolin et al. 1982; Voznesenskaya et al. 2008; Moir-Barnetson 2014).

214

215 **2.2 Analyses**

216 *2.2.1 Abundance and bulk stable isotopic values of total and organic sedimentary carbon*

217 As per reported in Rouillard et al. (2016), relative abundance (%) and stable isotope composition
218 (‰) of the sediment total C, organic C and total N were determined for every other interval (1 cm)

219 of core FOR106C3 using a Thermo Flash Elemental Analyser coupled with a Delta V Plus Isotope
220 Ratio Mass Spectrometer (Thermo Fisher Scientific, Bremen, Germany). Stable isotope results are
221 reported in permil (‰) after multi-point normalisation to international Vienna Pee Dee Belemnite
222 standard (VPDB) scale using international reference materials NBS19, USGS24, NBS30, LSVEC
223 (Skrzypek 2013), with combined analytical uncertainty <0.10 ‰. All samples were freeze-dried,
224 ground using mortar and pestle and homogenised prior to analyses. To isolate the organic
225 fraction, approximately 100–150 mg of ground sample was treated with 4 % HCl at room
226 temperature overnight to remove carbonates prior to analysis. Acid pre-treatment for the removal
227 of inorganic carbon (IC) can lead to biases beyond instrumental precision in C abundance and $\delta^{13}\text{C}$
228 measurements (e.g., 4 ‰; Brodie et al. 2011), which may become particularly significant when
229 total C-content is very low, as is the case with these Pilbara sediments (Table 2). The total C and
230 $\delta^{13}\text{C}$ results are therefore reported for total C without any acid pre-treatment i.e., without prior
231 removal of inorganic C in the sediments.

232

233 *2.2.2 Extraction and isolation of lipids*

234 The sediment record of Fortescue Marsh at 14 Mile Pool comprises four distinct stratigraphic
235 periods, together spanning several millennia of the marsh's environmental history (Rouillard et al.
236 2016). Replicate core FOR106C2 was separated into four sections each corresponding to one of
237 the four periods, which were then each ground using mortar and pestle, homogenised and
238 analysed for their molecular and stable carbon isotopic composition. Briefly, the sections and ages
239 of the four periods we investigated and described in Rouillard et al. (2016) corresponded to Period
240 1 (P1; 1–19 cm, CE 1990–2012); Period 2 (P2; 19–40 cm, CE 1600–1990); Period 3 (P3; 40–49 cm,
241 CE 700–1600) and 4 (P4; 49–61 cm, CE <100–700). Several dating methods (^{210}Pb , ^{137}Cs and ^{14}C)
242 have coherently estimated the age of the record to encompass the last ~2000 years. Details of the
243 dating procedures for the sediment record based on both replicates (FOR106C2 and FOR106C3)

244 and hydrological interpretations based on the analysis of lithology, moisture content, particle size,
245 microfossils, total C and N content and total stable isotope compositions ($\delta^{13}\text{C}$ and $\delta^{15}\text{N}$) on
246 replicate core FOR106C3 exclusively have been previously reported (Rouillard et al. 2016). To
247 summarise the hydrological interpretations, near perennial aquatic conditions were inferred for
248 P1; P2 sediments reflected a relatively drier and highly variable environment than P1 with episodic
249 extreme flood disturbance; P3 and P4 both reflected generally more arid conditions than P1 and
250 P2 and low energy hydrologies; conditions from the base of the record in P4 were likely hyperarid
251 until relatively less arid conditions and more regular flooding established during P3 from ~CE 700
252 (Rouillard et al. 2016). Unfortunately, concentrations of pollen grains in the sequence were very
253 low and high levels of deterioration were evident, likely reflecting the rewetting and subsequent
254 drying of the sediment cracking clays (Cushing, 1967; Lowe, 1982). Therefore, pollen could not be
255 used as a reliable indicator of past vegetation at this site (Rouillard et al. 2016).

256

257 The four samples (~20 g) from the FOR106C2 core were separately extracted in a Soxhlet
258 apparatus using a mixture of dichloromethane (DCM) and methanol (90:10, v/v) for 48-hr.
259 Elemental sulfur was removed through the extraction by the prior addition of activated copper
260 turnings to the sample (Blumer 1957). The total solvent-extracted bitumen was separated by small
261 column chromatography (activated silica stationary phase 150°C, 8h; 4 cm Pasteur pipette) into
262 saturate, aromatic and polar hydrocarbon fractions by successive elution with 2 mL *n*-hexane, 2
263 mL *n*-hexane:DCM (3:1) and 2 mL DCM:methanol (1:1) (Pagès et al. 2015). Fraction yields were
264 quantified but with relatively high uncertainty because of the very low masses extracted (total
265 extractable fraction <160 $\mu\text{g}\cdot\text{g}^{-1}$ dry sediment).

266

267 *2.2.3 Gas chromatography - mass spectrometry (GC-MS) analysis of hydrocarbon fractions*

268 GC-MS of the aliphatic and aromatic hydrocarbon fractions was conducted with an Agilent 6890
269 GC interfaced to a 5973 mass selective detector (MSD). The GC was fitted with a 60-m capillary
270 column (HP 5MS, cross-linked 5 % phenyl 95 % methyl-polysiloxane, 0.25 mm i.d., 0.25 µm film
271 thickness), used He carrier gas (1.1 mL·min⁻¹) and analyses were conducted in pulsed splitless
272 mode with the injector at 280°C. Full scan (*m/z* 50–550) 70 eV mass spectra were acquired, with
273 product assignments based on GC retention time or mass spectral correlation with the Wiley275
274 MS library, other laboratory standards or published data. Peak areas in the total ion
275 chromatogram (TIC) were used for quantification, except for steroids and terpenoids whose peak
276 areas were measured from *m/z* 217 and *m/z* 191 fragmentograms, respectively. Only products
277 detected at > 1 % the total TIC or fragmentogram signals were used for quantification.

278 The P_{aq} index was calculated for each sediment period. P_{aq} is defined as:

$$279 \quad (n-C_{23} + n-C_{25}) / (n-C_{23} + n-C_{25} + n-C_{29} + n-C_{31})$$

280 and has been used previously as a proxy for the relative contributions of algal, macrophytic and
281 terrestrial inputs in lacustrine environments (Ficken et al. 2000). In individual plants, P_{aq} values of
282 >0.4 are consistent with mainly algal input; 0.1–0.4 indicates a mainly macrophytic source and
283 <0.1 implies a mainly terrestrial origin.

284

285 *2.2.4 Compound-specific $\delta^{13}C$ analyses*

286 Compound-specific stable carbon isotope compositions were measured with an Agilent 6890 GC
287 coupled with a Micromass IsoPrime2 isotope ratio mass spectrometer (IRMS). The GC instrument
288 was fitted with a 60 m length, 0.25 mm i.d., 0.25 µm thick DB-1 phase column and analyses were
289 conducted in pulsed splitless mode following setup as in Pagès et al. (2015). Stable carbon isotope
290 compositions (‰), determined by integration of the *m/z* 44, 45 and 46 ion currents of the CO₂
291 peaks from each compound, were reported in the delta notation ($\delta^{13}C$) all reported values are
292 averages of at least triplicate analyses (standard deviation ≤0.2 ‰). The precision and accuracy of

293 the isotopic measurements were maintained by multi-point normalisation using the linear
294 regression slope and intercept obtained from frequent measurements ($n=38$; average standard
295 deviation = 0.2 ‰) of an in-house standard mixture of n -alkanes (n -C₁₄, -C₁₇, -C₁₉, and -C₂₅) of
296 known isotopic composition (respectively: -30.99, -29.72, -32.12, -31.94 ‰). Raw $\delta^{13}\text{C}$ values were
297 recalculated to VPDB scale using multipoint normalisation based on an in-house standard mixture
298 of four n -alkanes (n -C₁₄, -C₁₇, -C₁₉, and -C₂₅) of known isotopic composition (respectively: -30.99, -
299 29.72, -32.12, -31.94 ‰ VPDB). Analytical precision of the stable isotope measurements were
300 monitored by frequent measurements of laboratory standards ($n=38$), including the additional n -
301 C₁₁, n -C₁₃, and n -C₁₈ (respectively: -26.0; -29.3; -30.6 ‰ VPDB), during samples analyses and it was
302 ≤ 0.2 ‰ (one standard deviation).

303

304 2.2.5 Mixing model to determine sources of n -alkanes

305 Probability estimates of the proportional contribution of terrestrial C₃ versus C₄ arid plants to the
306 $\delta^{13}\text{C}$ values of the long-chain odd n -alkanes (n -C₂₇ to n -C₃₃) were obtained using the Bayesian
307 isotopic mixing model of the SIAR package (Parnell et al. 2010) in R v. 3.1.1 (MCMC = 30000). Two
308 source end-members in the mixing model were based on a subset of leaf wax $\delta^{13}\text{C}$ values from C₃
309 and C₄ plants exclusively from arid regions and with taxonomic relevance to the Upper Fortescue
310 River catchment (summarised in Table 1) to account foliar $\delta^{13}\text{C}$ differences caused by water stress
311 (Weiguo et al. 2005; Diefendorf et al. 2010; Kohn 2010; Schulze et al. 2014) and other adaptive
312 differences among species (e.g., Turner et al. 2008). Owing to the paucity of published data for
313 arid Australia, the source dataset included a subset of 50 species mainly from southern African
314 savannas. Published records of n -alkane $\delta^{13}\text{C}$ values (331 entries; 280 species) were restricted to
315 taxa (mainly Genera, but also Family) found in the eastern Pilbara (*FloraBase*;
316 www.florabase.dpaw.wa.gov.au).

317

319 **Table 1:** Relative abundances of *n*-alkanes (% odd *n*-C_{27–33}) and their stable isotopic signatures ($\delta^{13}\text{C}$)
 320 in leaves from C₃ and C₄ plant taxa found in the semi-arid Pilbara (Arid) and in various climates (All).
 321 Sources: Collister et al. 1994; Ficken et al. 2000; Chikaraishi and Naraoka 2003; Bi et al. 2005; Mead
 322 et al. 2005; Krull et al. 2006; Rommerskirchen et al. 2006; Tanner et al. 2007; Pedentchouk et al.
 323 2008; Vogts et al. 2009; Diefendorf et al. 2010; Kristen et al. 2010; Tanner et al. 2010; Diefendorf et
 324 al. 2011.

325

		<i>n</i> -C ₂₇	std	<i>n</i> -C ₂₉	std	<i>n</i> -C ₃₁	std	<i>n</i> -C ₃₃	std	<i>n</i>	mean	std
<u>Arid</u>												
Abundance	C ₃	21 ± 14		36 ± 19		30 ± 14		13 ± 15		20	25 ± 15	
	C ₄	11 ± 7		16 ± 7		45 ± 13		29 ± 15		30	25 ± 10	
$\delta^{13}\text{C}$	C ₃	-32.5 ± 2.1		-33.3 ± 1.8		-34.2 ± 2.2		-35.2 ± 1.9		8–22	-33.8 ± 2.0	
	C ₄	-22.0 ± 2.6		-21.9 ± 2.0		-22.4 ± 2.2		-22.3 ± 1.8		30–32	-22.2 ± 2.2	
<u>All</u>												
Abundance	C ₃	15 ± 17		38 ± 24		33 ± 19		14 ± 18		138	25 ± 19	
	C ₄	14 ± 12		18 ± 12		39 ± 15		29 ± 18		49	25 ± 14	
$\delta^{13}\text{C}$	C ₃	-32.4 ± 2.8		-36.2 ± 2.6		-33.7 ± 3.1		-36.8 ± 2.9		94–176	-34.7 ± 2.9	
	C ₄	-23.2 ± 5.1		-24.9 ± 5.0		-23.6 ± 5.1		-25.1 ± 5.5		67–88	-24.2 ± 5.2	

326

327 (2-COLUMN FITTING TABLE)

328

329 Descriptive statistics of the arid zone C₃ and C₄ end-members used in the mixing model ($\delta^{13}\text{C}$ and
 330 abundance) were then compared with those of the full dataset incorporating non-Pilbara
 331 vegetation to validate subsequent interpretation. As would be expected, *n*-alkanes of the
 332 terrestrial C₃ (*n* = 22) and C₄ (*n* = 32) plants formed distinct groupings within the $\delta^{13}\text{C}$ data set with
 333 minimum overlap (Table 1). Briefly, *n*-alkanes of C₃ plants had significantly (*p*<0.005) lower $\delta^{13}\text{C}$
 334 values than C₄ in both the arid exclusive (C₃=-33.8 ± 2.0 ‰; C₄=-22.2 ± 2.2 ‰) and the full plant
 335 datasets (C₃=-34.7 ± 2.9 ‰; C₄=-24.2 ± 5.2 ‰). The average $\delta^{13}\text{C}$ values for the *n*-alkanes of C₃
 336 and C₄ plants were both slightly higher and had smaller standard deviations in the arid dataset
 337 than the full dataset; however, this difference was not significant (*p*>0.1). Based on the same
 338 subset, we used an abundance matrix (Table 1) of the selected *n*-alkanes relative to one another
 339 (%) to account for the effect of concentration-dependence biases on mixing model estimates
 340 (Phillips and Koch 2002). The average *n*-alkane abundances calculated for C₃ or C₄ sources from

341 the full datasets as well as the smaller subset of arid zone data were also consistently and
342 significantly different (p -value <0.005), except for n -C₂₇ ($p=0.7$) and n -C₃₁ ($p=0.06$). In addition, n -
343 C_{27, 29} were more abundant in C₃ plants while n -C_{31, 33} were more abundant in C₄ plants, and was
344 most pronounced for arid species. However, there was no significant difference ($p=0.06$) in the
345 relative abundance of the n -alkanes between the full and arid only datasets for either C₃ or C₄.

346

347

348 **3. RESULTS**

349 **3.1 Organic matter composition**

350 *3.1.1 Bulk OC and solvent-extractable compounds*

351 Sediment C and N abundances and stable isotope compositions are described in relation to each
352 of the four hydroclimate periods described previously for this region: P1 (1–19 cm, CE 1990–2012),
353 P2 (19–40 cm, CE 1600–1990), P3 (40–49 cm, CE 700–1600) and P4 (49–61 cm, CE <100–700).
354 Total C was very low in all sediments (Table 2). TC content was highest in P1 (1.4 %) and lowest in
355 P4 (< 0.4 %), with P2 (0.6 %) and P3 (0.7 %) between these values. Similarly, the highest N content
356 as well as lowest $\delta^{15}\text{N}$ were found in P1 (0.12 % wt and 6.1 ‰). Conversely, the lowest N content
357 and highest $\delta^{15}\text{N}$ were found in P4 (0.01 % wt and 7.6 ‰). P2-P3 showed similar intermediate
358 values (0.03 % wt and 7.4-7.5 ‰). The total solvent-extractable bitumen fraction (Table 2) of P1–
359 P3 was 60 to 150 $\mu\text{g}\cdot\text{g}^{-1}$ of sediment, representing around 1 % of total C. Polar compounds
360 dominated the bitumen (NSO - nitrogen, sulfur and oxygen polar compounds - fraction >89 %),
361 whereas the aliphatic fractions were between 2 and 5 % (P3>P1>P2) and the aromatic fraction
362 even smaller in P1 and P3 (<1 %). The aromatic fraction of P2 was relatively high at 8.5 %. The
363 extractable fraction (i.e., bitumen) of P4 sediment was <5 $\mu\text{g}\cdot\text{g}^{-1}$ of sediment, which was
364 insufficient for further characterisation. Sediment $\delta^{13}\text{C}$ values for the entire core were in the range
365 -20.3 to -28.2 ‰, which are typical values of biologically sourced sedimentary OM (e.g., with

366 mixed C₃/C₄ inputs) and imply negligible carbonate influence on measured $\delta^{13}\text{C}$ (Finlay and
 367 Kendall 2008).

368

369 **Table 2:** Geochemical parameters of four organic sedimentary periods (P1–P4) from the 14 Mile
 370 Pool sediment record, NW Australia.

371

Parameter		P1 (1990–2012 ¹)	P2 (1600–1990)	P3 (700–1600)	P4 (<100–700)	Total Av (<100–2012)
Bulk sediment content & stable isotopes	n (combined intervals)	17	23	7	11	58
	Total C (%wt)	1.4	0.6	0.7	<0.4	0.8
	Range	0.8–5.6	<0.4–1.1	0.5–0.9	<0.4–0.5	<0.4–5.6
	Total N (%wt)	0.12	0.03	0.03	0.01	0.05
	Range	0.06–0.17	0.01–0.06	0.02–0.03	<0.01–0.01	<0.01–0.17
	$\delta^{13}\text{C}$ [‰ VPDB]	-24.3	-22.9	-20.3	-28.2	-23.4
	Range	-26.7–-22.3	-26.2–-20.2	-22.4–-16.9	-28.4–-28.0	-28.4–-16.9
	$\delta^{15}\text{N}$ [‰ AIR]	6.1	7.5	7.4	7.6	7.1
	Range	4.8–8.4	6.5–7.9	7.0–7.6	7.5–7.8	4.8–8.4
	Yield					
	Tot ext ($\mu\text{g}\cdot\text{g}^{-1}$)	155	87	63	< 5	-
	Aliphatic (% tot)	4.7	2.6	5.0	-	-
	Aromatic (% tot)	0.3	8.5	0.7	-	-
	NSO (% tot)	95	89	94	-	-
Aliphatic fraction	Alkanes (%TIC)	84	61	71	-	-
	Steroids (%TIC)	1.4	8.1	7.2	-	-
	Hopanoids (%TIC)	2.7	4.2	5.8	-	-
	Others (% TIC)	12	27	16	-	-
CPI	$n\text{-C}_{23-33}$	3.1	5.7	5.0	-	-
P _{aq}		0.3	0.1	0.3	-	-
Mixing	C ₃ (%)	59	53	50	-	-
	C ₄ (%)	41	47	50	-	-

372 **Note:** ¹All time periods are provided in years of the Common Era (CE); C and N content (Total C and Total N) and stable carbon and
 373 nitrogen isotopes ($\delta^{13}\text{C}$ and $\delta^{15}\text{N}$) (modified from Rouillard et al. 2016); estimated yields of the total extractable fraction (Tot ext; $\mu\text{g}\cdot\text{g}^{-1}$
 374 of sediment); the aliphatic, aromatic and polar (NSO) fractions are expressed as a % of the total extractable fraction; relative abundance
 375 of *n*-alkanes, steroids, hopanoids and other aliphatic compounds on the total ions chromatogram (%TIC); carbon preference index (CPI)
 376 and P_{aq} index (see
 377

378 (2-COLUMN FITTING TABLE)

379

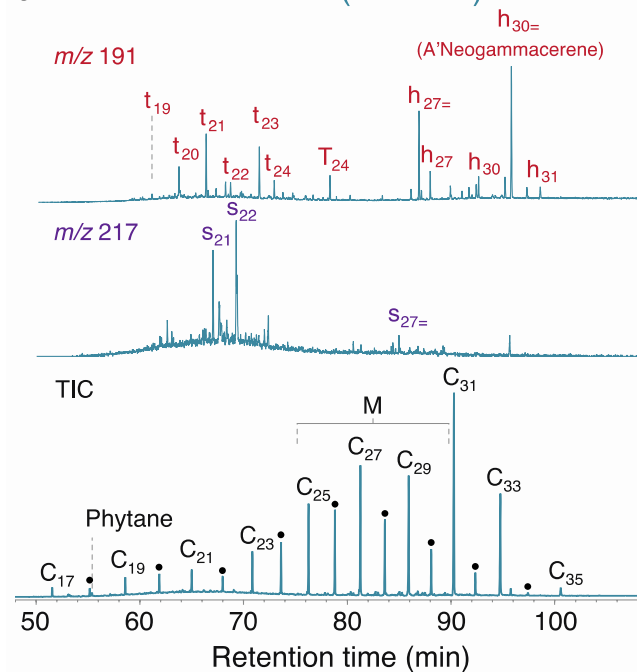
380 3.1.2 Aliphatic hydrocarbon fractions

381 GC-MS analysis of the aromatic fractions yielded very few products; however, a wide distribution
 382 of aliphatic hydrocarbons were detected in the saturated fractions of P1–P3. The main aliphatic
 383 products included: *n*-alkanes, regular isoprenoids (e.g., phytane), branched alkanes and alkenes
 384 and saturated and unsaturated hopanoids and steroids. A particularly prominent homologous
 385 series of C₁₇ to C₃₅ *n*-alkanes accounted for ~60 to 85 % of the TIC signal of the saturate fraction,
 386 where *n*-C₃₁ was the most dominant analyte (15–21 % of TIC; Fig. 2). The carbon preference index
 387 (CPI) was separately calculated for long-chain (C₂₃ to C₃₃) *n*-alkanes (e.g., Bray and Evans 1961;

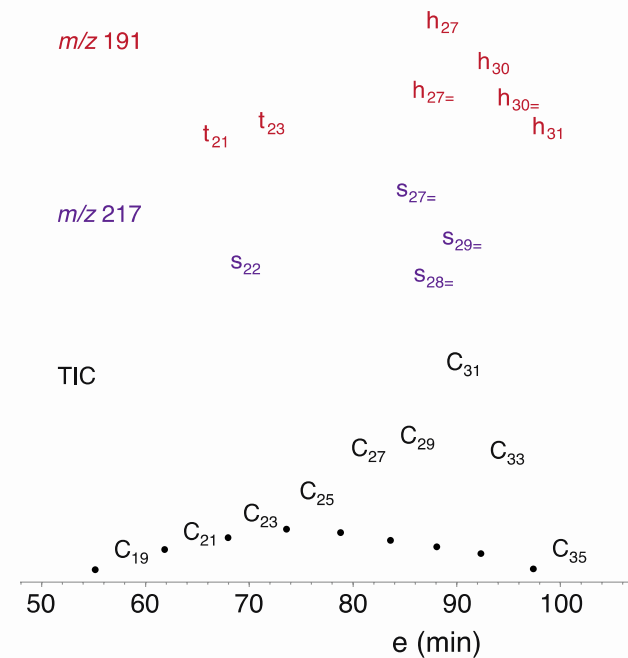
388 Kuhn et al. 2010; Rao et al. 2008). The longer chain members showed a strong odd-over-even
389 predominance ($CPI_{23-33} = 3-6$; $P2 > P3 > P1$; Table 2). The P_{aq} values of P1–P3 were 0.1 or 0.3
390 suggesting the OM may mostly originate from macrophytes (Ficken et al. 2000). However, it is
391 important to note that the P_{aq} value in sediment extracts likely reflect a mixture of inputs and may
392 therefore also integrate a signal from both terrestrial and aquatic sources (Ficken et al. 2000). Less
393 abundant products across P1–P3 included a similar series of monounsaturated *n*-alkenes, as well
394 as several branched alkanes including the regular isoprenoids pristane and phytane and polycyclic
395 hopanoids and steroids. The hopanoids, best seen from the *m/z* 191 fragmentogram (Fig. 2 top),
396 included C_{27} to C_{31} hopanes and even more abundant mono- and diunsaturated hopenes ($hC_{28-30=}$),
397 tricyclic terpanes (tC_{19-24}) and the C_{24} tetracyclic terpane (TC_{24}). The steroids included C_{28} to C_{30}
398 steranes (sC_{28-30}) as well as several short chain steranes (sC_{19-22}). The distribution of aliphatic
399 products detected across sediment depths (P1–P3) were broadly similar. However, pristane and
400 phytane were only detected in P1. The P1 sediments also showed a greater diversity, but lowest
401 relative abundance, of hopanoid and steroid products. In contrast, *n*-alkenes were detected in P2
402 and P3 but not P1. Sediments of P2 had the highest representation of aliphatics other than *n*-
403 alkanes. These were mainly linear branched alkanes and alkenes; short chain steranes (sC_{19-22})
404 were the major steroids of P1 (>50 % of *m/z* 217 steroid signal) while sterenes were particularly
405 high in P2 and P3 (>80 % of *m/z* 217; Fig. 3).

406

a.P1: 1990-2012 CE (0-20 cm)



c.P3: 700-1600 CE (42-48 cm)



407

408

409 **Figure 2:** GC-MS total ion chromatograms (TIC) and fragmentograms (m/z 191 and 217) of the three periods in the core; (a) P1, (b) P2 and (c) P3.

410 Tentatively assigned compounds are labelled with their corresponding carbon number and unsaturation level (=); odd (C) and even (filled circles) n -

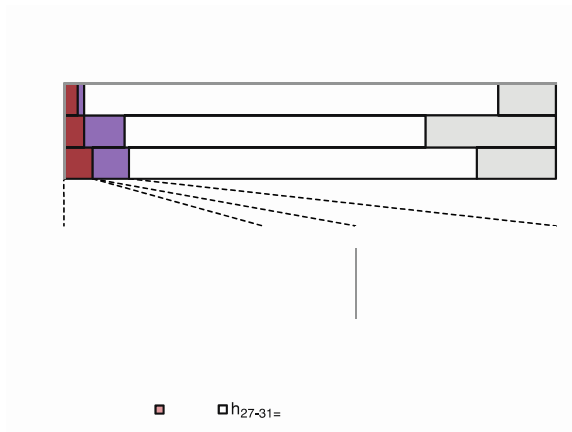
411 alkanes, tricyclic terpanes (t), tetracyclic terpanes (T), steroids (s), hopanoids (h) methylated alkane series range (M) and linear aliphatic series (open

412 circles).

413 (1.5-COLUMN FITTING IMAGE)

414

415



416

417 **Figure 3:** Relative peak abundances (% TIC) of the main aliphatic compounds extracted by GC-MS
418 and quantitative distributions of hopanoids and steroids within each sedimentary period.

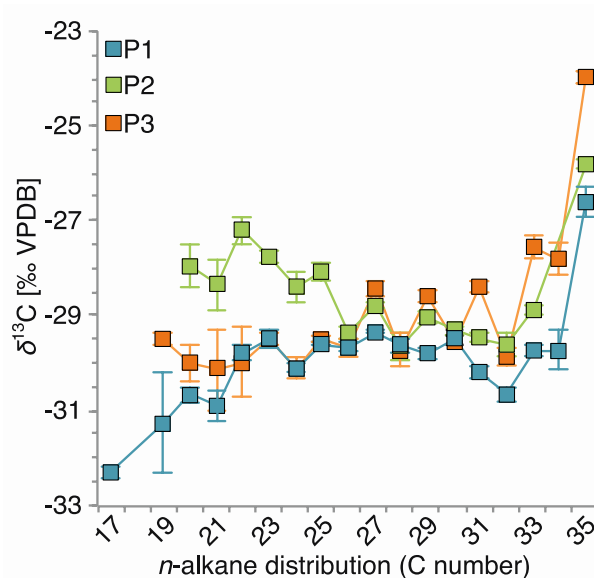
419 (1-COLUMN FITTING IMAGE)

420

421 3.1.3 $\delta^{13}\text{C}$ of *n*-alkanes

422 Stable carbon isotope values were measured for most *n*-alkanes detected in the P1–P3 samples
423 (Table S1; Fig. 4). However, the concentrations of other aliphatic products were too low for
424 accurate $\delta^{13}\text{C}$ measurement. The $\delta^{13}\text{C}$ value range for *n*-alkanes was mostly in a relatively narrow
425 range (± 2 ‰) around -29 ‰. Overall minimum $\delta^{13}\text{C}$ was -32.3 ± 0.1 ‰ (*n*-C₁₇ in P1) and maximum
426 $\delta^{13}\text{C}$ was -24.0 ‰ (*n*-C₃₅ in P3). The $\delta^{13}\text{C}$ values of the mid-chain *n*-alkanes (*n*-C_{21–26}) were similar
427 in P1 and P3, but higher by ~ 2 ‰ in P2. The odd *n*-alkanes had higher $\delta^{13}\text{C}$ values compared to
428 their even carbon neighbours, particularly for the longer-chain compounds ($>n\text{-C}_{27}$) where this
429 differential was as much as ~ 1 ‰ in sediments from for P2. The $\delta^{13}\text{C}$ values of the odd numbered
430 *n*-alkanes from *n*-C₂₇ followed the trend of P1 < P2 < P3 i.e., increasing with sediment depth and age
431 (Fig. 4).

432



433

434

435 **Figure 4:** Average carbon stable isotopic signature ($\delta^{13}\text{C}$) of the *n*-alkanes extracted from each
 436 sediment period: P1 (blue), P2 (green) and P3 (orange). Error bars represent the standard
 437 deviation obtained from triplicate measurements.

438 (1-COLUMN FITTING IMAGE)

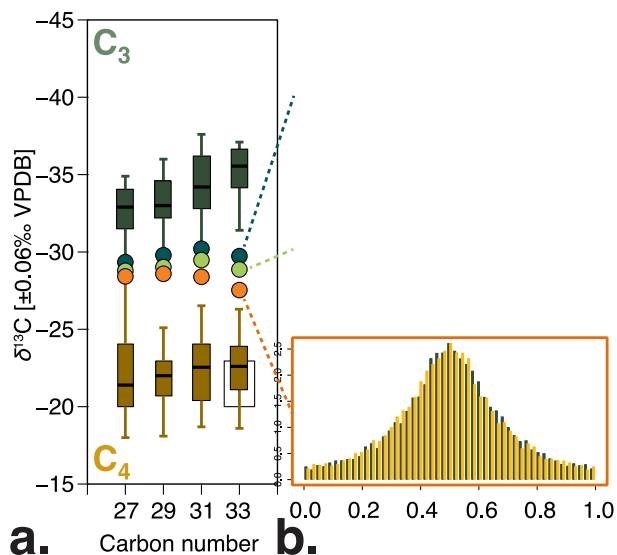
439

440 3.2 Sources of organic matter based on mixing model of long-chained *n*-alkanes

441 The mixing model based on $\delta^{13}\text{C}$ values of the long-chain odd *n*-alkanes (*n*-C₂₇ to *n*-C₃₃) for P1–P3
 442 is shown in Figure 5b. Typical of two-end member mixing models all probability distributions were
 443 unimodal (Parnell et al. 2010; Hopkians and Ferguson 2012; Fig. 5). The model suggests terrestrial
 444 C₃ dominance over recent decades (~20 years) (P1 median = 59 %; Table 2; Fig. 5), but with
 445 increasing importance of C₄ OM sources in the sediments of P2 (C₃ median = 53 %) to co-
 446 dominance of C₃ and C₄ sources in P3 (C₃ ~ C₄ = 50 %). The probability distributions for C₃ and C₄
 447 are wide and overlapping indicating a high uncertainty with these estimates, however diverging
 448 median values suggests an increasing C₃ contribution and corresponding decrease in the C₄ with
 449 time (Fig. 5). While small aquatic and macrophytic inputs were detected in the molecular
 450 distributions, their contributions are unlikely to weight significantly on the mixing model estimates

451 given the strong terrestrial plant signal of the abundant long-chain *n*-alkanes used here (see
452 Discussion below).

453



454

455

456 **Figure 5: a)** Carbon stable isotopic signatures ($\delta^{13}\text{C}$) of long chain *n*-alkanes C₂₇, C₂₉, C₃₁ and C₃₃

457 extracted from the P1 to P3 sediments: P1 (dark blue circles), P2 (light green circles) and P3

458 (orange circles) compared to boxplots showing the range of terrestrial leaf waxes values of C₃

459 (light and dark green) and C₄ (light and dark yellow) plants as described in the literature (wide light

460 coloured boxes) and the arid zone subset (thin dark coloured boxes); **b)** change in end-members

461 contribution over time, i.e., probability distributions of contribution from C₃ (dark green) and C₄

462 (yellow) sources (SIAR).

463 (1-COLUMN FITTING IMAGE)

464

465

466 4. DISCUSSION

467 4.1 The organic molecular and stable isotopic fingerprint of drylands

468 These results show that it is possible to extract organic molecular fossil distributions and $\delta^{13}\text{C}$ of *n*-
469 alkanes from lake sediments of low OM content in tropical arid ecosystems. The low total C
470 content of the Fortescue Marsh sediments (<1 %) reflects the inherently low OM contents of
471 modern Pilbara soils and is typical of dry stream-bed sediments (Bennett and Adams 1999; Ford et
472 al. 2007; McIntyre et al. 2009b). The low concentration of the extracted bitumens (60–150 $\mu\text{g}\cdot\text{g}^{-1}$
473 of bulk P1–P3 sediment; Table 2) is also typical of OM-poor environments and oligotrophic aquatic
474 systems (e.g., Kristen et al. 2010; Pautler et al. 2010; Fang et al. 2014). Sub-aerial exposition of the
475 sediment during prolonged dry periods over recent millennia (Rouillard et al. 2016) would also
476 have reduced OM preservation (Verschuren 1999). Nevertheless, saturated hydrocarbons typical
477 of lipid metabolites were found in the upper 49 cm of the core (P1–P3; Fig. 2).

478

479 Our results also show that the molecular fossils can be used to characterise the major sources of
480 sedimentary OM from the tropical arid system. The major aliphatic products identified in our
481 sediments are indicative of predominantly terrestrial sources of OM at 14 Mile Pool (Fig. 2). The
482 sediment of P1–3 consistently showed a homologous series of *n*-alkanes peaking at *n*-C₃₁. *n*-
483 Alkanes are common hydrocarbon products of organic lake sediments derived from fatty acids,
484 which are abundant components of epicuticular leaf waxes (Eglinton and Hamilton 1967) as well
485 as many other biological systems and so are not always particularly source diagnostic of exact
486 sources without additional evidence (e.g., stable isotopic composition). The relatively high
487 abundances of *n*-alkanes in sediments across ecosystems are also due to a high structural
488 recalcitrance to diagenesis (i.e., relatively high chemical stability, low solubility and are degraded
489 slowly by microbes) compared to other functionalised biochemicals (Cranwell 1981; Schmidt et al.
490 2011). The unique lipid composition of many biological systems gives rise to *n*-alkanes of
491 distinctive molecular distributions or stable isotope patterns that have thus been used previously
492 to infer past hydroclimatic conditions across a range of environments (Castañeda and Schouten

493 2011; Garcin et al. 2012; Leng and Henderson 2013; Sun et al. 2013). For instance, chain length
494 distribution can be used to distinguish algal (short chain) from terrestrial, higher plant (long chain)
495 sources (Gelpi et al. 1970; Meyers and Ishiwatari 1993). A relatively high abundance of long chain
496 *n*-alkanes ($>n\text{-C}_{25}$) is often attributed to a high flux (or persistence) of terrestrially-derived OM
497 (Gelpi et al. 1970; Meyers and Ishiwatari 1993; Freeman and Colarusso 2001). While there are
498 non-terrestrial sources of long chain *n*-alkanes e.g., *Botryococcus braunii*, a phytoplanktonic alga
499 that is common in hypersaline or evaporitic lake settings (Grice et al. 1998), the long chain *n*-
500 alkanes prominent in each of the P1–P3 sediments also showed a distinctive odd/even carbon
501 number preference ($\text{CPI}_{23-33} = 3-6$; Fig. 2) typical of terrestrial plant sources (allochthonous; e.g.,
502 Meyers 1997; Volkman et al. 1998; Freeman and Colarusso 2001; Rao et al. 2008).

503

504 The *n*-alkanes detected in the P1–P3 sediments ($n\text{-C}_{15-35}$ and $\text{CPI}=2-12$; Fig. 2 and Tables 2 & S1)
505 had a similar distribution to those reported in many other arid zone lake deposits (e.g., Garcin et
506 al. 2012). The $\delta^{13}\text{C}$ values were in the range -31 to -24 ‰ and displayed a ‘saw-toothed’ odd
507 versus even carbon numbered pattern (Fig. 4) that is very typical of a leaf wax source and linked
508 with higher plants biosynthesis (Collister et al. 1994; Zhou et al. 2010). The ephemeral reaches of
509 low-lying rivers are often important sinks for the terrestrial OM of upper catchment rivers,
510 streams and pools in arid environments (Jacobson et al. 2000; Fellman et al. 2013). Allochthonous
511 material transported from the catchment through floods or wind is one of the main drivers of
512 ecosystem metabolism in river networks (Sponseller et al. 2013). $\delta^{13}\text{C}$ of the 14 Mile Pool *n*-
513 alkanes (-32.3 to -26.6 ‰ Fig. 4) are consistent with observations from modern sediments (-35.9
514 to -26.8 ‰) from the arid, Sudano-Sahelian shrub savannah zone of Cameroon (Garcin et al.
515 2012). The CPI values of long-chain *n*-alkanes from dryland soils can vary with aridity (Krull et al.
516 2006; Rao et al. 2008; Luo et al. 2012) and can extend to higher values than those reported here
517 (e.g., 3–13, for $n\text{-C}_{23}\text{--}n\text{-C}_{34}$ in several Chinese soils; Rao et al. 2008). Consequently, the molecular

518 evidence from the 14 Mile Pool sediments records a clear and dominant signature of the
519 terrestrial plant inputs from the surrounding arid catchment.

520

521 The shorter chain *n*-alkanes (*n*-C_{17–26}) of the sediments from 14 Mile Pool had lower $\delta^{13}\text{C}$ values
522 than the more abundant long-chain *n*-alkanes (*n*-C_{27–35}), implying that these *n*-alkanes derive from
523 a different source. The $\delta^{13}\text{C}$ of low MW *n*-alkanes (<*n*-C₂₅) in P1 and P3 were up to 1.5 ‰ lower
524 than co-occurring odd numbered high MW *n*-alkanes (>*n*-C₂₇; Fig. 4). These lower $\delta^{13}\text{C}$ values are
525 consistent with the lipids of algae or aquatic plants (Gelpi et al. 1970; Ficken et al. 2000), which
526 typically have lower $\delta^{13}\text{C}$ values than plant lipids (Chikaraishi and Naraoka 2003). Lower MW *n*-
527 alkanes typically derive from lower trophic organisms and aquatic producers such as
528 phytoplankton, benthic algae or aquatic macrophytes (e.g., Gelpi et al. 1970; Mead et al. 2005;
529 Aichner et al. 2010; Garcin et al. 2012). Macrophyte input was also indicated by P_{aq} values of 0.1–
530 0.3 for P1 and P3 sediments, though a mixture of the terrestrial and aquatic sources may also have
531 contributed to this value (Ficken et al. 2000). In the P2 sediments, the P_{aq} value was in the same
532 range (0.3) but the short chain *n*-alkanes were ~ 2 ‰ more positive than the long chain *n*-alkanes.
533 The $\delta^{13}\text{C}$ values of the P2 sediments were also similar to the $\delta^{13}\text{C}$ values of *n*-alkanes reported for
534 Lake Koucha sediments attributed to inputs from macrophytes and algae that had likely been
535 influenced by higher $\delta^{13}\text{C}$ of bicarbonates in the water body (e.g., Aichner et al. 2010; Kristen et al.
536 2010). Speculatively, the heavier short-chain *n*-alkanes in P2 may suggest a relatively stronger
537 groundwater influence during this period on the pool's hydrology and chemistry than present. This
538 observation would be consistent with the occurrence of one or several extreme floods inferred for
539 the periods (Rouillard et al. 2016), which would have enable an important rise in the water table
540 (Skrzypek et al. 2013) and favourable conditions for macrophytic growth, i.e. prolonged inundation
541 an a clear water column.

542

543 4.2 *n*-alkanes as proxies of ecosystem carbon stocks

544 By integrating the molecular abundance data with the Bayesian modelling of the stable isotopic
545 composition of the *n*-alkanes, we can assess that our results reflect the major vegetation types in
546 the catchment (C_3 and C_4) and hydrological processes at present and with hydroclimatic change
547 through the late Holocene in this subtropical arid zone. The particularly high abundance of *n*- C_{31} in
548 the sediments from 14 Mile Pool (Fig. 2) is consistent with the presently high catchment coverage
549 and biomass of warm-climate adapted C_4 grasses (~64 % of mainly *Triodia* spp.) relative to that of
550 woody C_3 plants, predominantly *Acacia* trees and shrubs (Fig. 1b; Adams et al. 2001). Though the
551 abundance of *n*- C_{31} alkane relative to *n*- C_{27} and *n*- C_{29} alkanes have been widely used to
552 differentiate grass-type (C_4) versus woody (C_3) plant contributions, recent research has shown that
553 caution is required with these estimates because the relative abundances of high molecular
554 weight *n*-alkanes can vary for both functional types (Bush and McInerney 2013). At 14 Mile Pool,
555 the occurrence of the very long-chain C_{33} (-26.6 ‰) and C_{35} (-24.0 ‰) *n*-alkanes and their
556 relatively high $\delta^{13}C$ signatures (Cf. Ave ~-28.5 ‰ of *n*- C_{33-35}); Fig. 4) are further evidence of C_4
557 plant input to the P1-3 sediments. Very long-chain *n*-alkanes are not only generally more
558 abundant in C_4 plants (e.g., Table 1), but their production may also be further enhanced under hot
559 and dry conditions (Rommerskirchen et al. 2006; Vogts et al. 2009). The $\delta^{13}C$ of *n*-alkanes in
560 Australian grassland soils usually reflects a C_4 prevalence (e.g., Krull et al. 2006).

561

562 The quantitative estimates from the stable isotope data suggests, however, the C_4 plants (41–50 %
563 median contribution in mixing model) might not have contributed as considerably to the 14 Mile
564 Pool sediments as might have been inferred from their current landscape dominance (64 %
565 aboveground biomass; Table 2; Fig. 5). This is because the riparian vegetation of the 14 Mile Pool
566 and broader riverine network (Fig. 1) may contribute substantial C_3 biomass resulting in a greater
567 contribution of ^{13}C -depleted *n*-alkanes than might be expected from the non-riparian C_3 plant

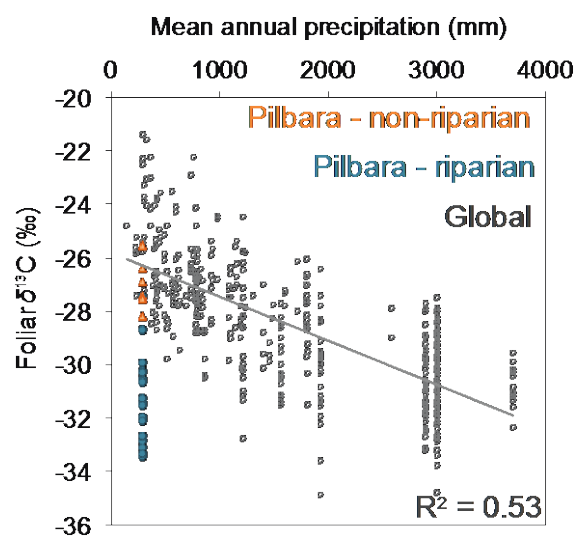
568 sources that predominantly inhabit arid environments. Disproportionally high riparian
569 contributions to organic sediment loads have previously been reported in other grasslands
570 dominated catchments: riparian trees in grasslands of sub-tropical Madagascar were thought
571 responsible for sedimentary OM with $\delta^{13}\text{C}$ values more than 6 ‰ lower than predicted from the
572 catchment vegetation composition (Marwick et al. 2014). Similarly, $\delta^{13}\text{C}$ trends in Holocene
573 sediments from a stream channel in southeastern Arizona were attributed to the establishment of
574 a riparian fringe and increasing contribution of C_3 herbaceous and woody vegetation over the
575 modern dominance of C_4 perennial grasses (Biedenbender et al. 2004). Preferential deposition of
576 riparian tree leaves to lake sediment and the importance of lateral OM contribution to streams
577 and rivers have also long been recognised in temperate environments (e.g., Connors and Naiman
578 1984; Seki et al. 2010). Viewed collectively, these findings suggest that the spatial structure of
579 catchment vegetation, rather than the cover or biomass only, strongly affects terrestrial OM
580 inputs to dryland sediments, and that the riparian contribution should particularly be taken into
581 consideration when interpreting vegetation changes in arid environments from sediment *n*-
582 alkanes $\delta^{13}\text{C}$.

583

584 If riparian plants (mainly C_3 and usually lower $\delta^{13}\text{C}$) are used as source material in the mixing
585 model, however, this would also likely change the $\delta^{13}\text{C}$ values of the C_3 end member, resulting in a
586 potential higher estimation of C_4 contribution to the Fortescue Marsh (Fig. 5). Response to plant
587 water stress, observed both globally (Weiguo et al. 2005; Diefendorf et al. 2010; Kohn 2010) and
588 along regional gradients (Schulze et al. 2014), can lead to an important $\delta^{13}\text{C}$ differential between
589 riparian and to non-riparian C_3 sources (Fig. 5). A summary of $\delta^{13}\text{C}$ foliar signatures from woody C_3
590 plants of this study's region showed that species from riparian fringes in the Pilbara have
591 significantly lower values (-4 ‰) than the foliage of homologue populations located on higher,
592 more water-stressed, parts of the landscape (Table S2; Fig. 6). Thus, C_3 -sourced *n*-alkanes with

593 $\delta^{13}\text{C}$ values 4 ‰ lower than non-riparian sources would significantly decrease the model estimates
 594 for the C_3 contribution and correspondingly project a higher C_4 input. A further complication to
 595 consider is the substantial $\delta^{13}\text{C}$ offset that can occur between *n*-alkanes, leaf lipids and whole
 596 leaves (Rieley et al. 1993; Bi et al. 2005; Diefendorf et al. 2011; Codron et al. 2013). The $\delta^{13}\text{C}$
 597 dynamics related to water availability such as local growth conditions (Schouten et al. 1998;
 598 Riebesell et al. 2000) and seasonal fluctuations (Lockheart et al. 1997) might also diverge for these
 599 entities, and there might also be variation of behaviour within phenology (Diefendorf et al. 2011),
 600 plant functional type (Shepherd and Wynne 2006; Bush and McInerney 2013; Cernusak et al.
 601 2013), leaf ontogeny (Nguyen Tu et al. 2004) and phylogenetic group (Pedentchouk et al. 2008;
 602 Diefendorf et al. 2011).

603



604

605

606 **Figure 6:** Relationship between foliar $\delta^{13}\text{C}$ and mean annual precipitation of woody (C_3) riparian
 607 (filled blue circles) and non-riparian (filled orange triangles) vegetation from the Pilbara (collected
 608 between 1997 and 2013 by the Ecosystem Research Group at The University of Western Australia;
 609 see Table S2 for more information of species and sampling of the Pilbara dataset) compared to the
 610 global distribution (open grey circles; modified from Diefendorf et al. 2010).

611 (1-COLUMN FITTING IMAGE)

612

613 In low water flow systems of arid zones, additional consideration of the complex deposition
614 processes across timescales may be important from the reconstruction of past catchment
615 vegetation from sediment records. For instance, leaf litter may accumulate in isolated pools or
616 onto dry stream beds (McIntyre et al. 2009b) until transportation in various stages of decay
617 downstream by heavy and unpredictable summer floods (e.g., Robertson et al. 2001; Francis and
618 Sheldon 2002; Mladenov et al. 2007). Ecological 'booms' following large rains contrast prolonged
619 periods of low productivity in arid catchments (Collins et al. 2008; McGrath et al. 2012; Kanniah et
620 al. 2013). High energy floods in arid regions can lead to a wide dispersal of higher plants across the
621 landscape and the increased likelihood following mortality of rapid photodegradation (i.e.,
622 minimal tree canopy or litter shelter from sustained UV radiation) or consumption by the termites
623 prolific in Australian savannahs (e.g., Holt et al. 1987; Chen et al. 2003; Austin et al. 2004). The
624 development of complementary indicators and of additional sediment records with finer temporal
625 resolutions from the arid zone sediment may help towards a better understanding of these
626 complex spatio-temporal patterns of OM deposition. Clearly, the range of $\delta^{13}\text{C}$ values for *n*-
627 alkanes from different sources needs to be better defined in space and time for more precise
628 quantitative mixing model estimates of contributions of different lipid sources to lake
629 environments and reconstruction of vegetation cover in catchments.

630

631 The molecular data also showed a small but distinctive aquatic contribution in P1 that was not
632 detected in older sediments (Fig. 2), particularly as determined from the lower chain length *n*-
633 alkanes (Gelpi et al. 1970; Mead et al. 2005; Aichner et al. 2010; Garcin et al. 2012). The
634 significantly lower $\delta^{13}\text{C}$ value of *n*-C₁₇ compared to the larger *n*-alkanes (Fig. 4) further suggests a
635 cyanobacterial contribution to sediment OM (Kristen et al. 2010). Cyanobacterial mats are
636 common in freshwater/brackish limnological conditions and high pH environments (Rontani and

637 Volkman 2005) and microbial mats were visually evident at the sediment/water interface of the
638 pool during several field surveys.

639

640 Steroids, generally originating from eukaryotic organisms such as higher plants or algae, as well as
641 hopanoids, which are typically associated with microbial inputs (Aichner et al. 2010), were
642 detected in all sediments. Their relative abundances were generally low (Fig. 3), though higher in
643 P2 and P3 (steroids 7–8 % TIC signal; hopanoids 4–6 %) than P1 (steroids 1.3 %; hopanoids 2.6 %),
644 perhaps on account of greater diagenetic reduction of the poly-hydroxy functionalities of their
645 biochemical form (Peters et al. 2005). Sediments of P1 showed a more diverse assemblage of
646 these products (Fig. 2 and 3), including unusual series of low molecular weight steranes (C_{19–22})
647 and tri- and tetracyclic terpanes (C_{19–C₂₄}). The greater biomarker diversity evident in P1 suggests
648 the recent wetter conditions (Rouillard et al. 2016) favoured a more complex microbial pool that
649 may also have supported greater heterotroph reworking of the OM (i.e., resulting in overall lower
650 abundance). A similar series of low molecular weight steranes (extending to slightly large C_n) were
651 reported in relatively high concentrations in immature evaporitic marl sediments of the Jinxian
652 Sag, Bohai Bay Basin, North China (Lu et al. 2009) and generally attributed to early diagenetic
653 products of microbial sources in a carbonate environment (Erik 2011; Oliveira et al. 2012). The
654 higher steroid to hopanoid ratio detected in P2 and P3 (Fig. 3) is also consistent with a greater
655 representation of terrestrially-derived OM during those periods compared to the more recent P1
656 sediments (Tissot and Welte 1984).

657

658 **4.3 Do sediment *n*-alkanes reflect hydroclimatic change at the Fortescue Marsh in the late** 659 **Holocene?**

660 The final objective of this study was to assess whether the distribution of biomarkers and $\delta^{13}\text{C}$
661 values of *n*-alkanes from the sediment reflected ecosystem responses to regional hydroclimatic

662 change over the late Holocene (last ~2000 years). While we previously discussed our molecular
663 record against the hydroclimatic reconstruction at the Fortescue Marsh (Rouillard et al., 2016),
664 here we also compare our findings to other regional palaeoecological records focussing on OM
665 and vegetation changes for the four periods investigated (P1-P4). The very low organic content of
666 the older section (P4) core—below the detection limit of presently applied GC-MS analyses—is
667 consistent with notably drier conditions in the inland Pilbara region more than 1300 years ago
668 compared to present (Rouillard et al. 2016). Marine sediment pollen records (van der Kaars et al.
669 2006) and coastal speleothem records (Denniston et al. 2013) have revealed a decrease in C₃
670 vegetation across northwest Australia since the mid-Holocene, which has been attributed to a
671 drier climate, equivalent to a decrease of ~150 mm·year⁻¹ in mean coastal rainfall. Terrestrial
672 pollen records further east also suggest a decline in effective precipitation (EP) in tropical northern
673 Australia (Gulf of Carpentaria) from 4000–3500 years BP followed by an increase in EP over the
674 last 2000 years (Schulmeister and Lees 1995). Recent evaluation of a sediment record from the
675 eastern Kimberley has also indicated increasing woody vegetation following rapid increased fluvial
676 sedimentation around CE 650–850 (McGowan et al. 2012).

677

678 Given the differences evident in the molecular and stable isotopic distributions of lipid derived *n*-
679 alkanes in sediments from the oldest (P3) to the youngest (P1), we might conclude that vegetation
680 structure and composition in the catchment have been broadly consistent in the region for the last
681 ~1300 years (Table 2; Fig. 2). Nonetheless, other hydroclimatic reconstructions based on physical
682 and other biogeochemical characteristics indicate conditions in the region have become
683 progressively wetter since ~ CE 700 (Rouillard et al. 2016 and references therein), including the
684 more coarsely resolved studies mentioned earlier. In particular, the near-permanently inundated
685 conditions under a wetter recent climate in the last ~20 years inferred from multiple sediment
686 indicators at the site (P1; Rouillard et al. 2016) are corroborated by reconstructions from tree-ring

687 and stable isotopes chronologies in the catchment (Cullen and Grierson 2007; O'Donnell et al.
688 2015), and by instrumental data and modelling from satellite imagery and rainfall timeseries
689 (Rouillard et al. 2015; Shi et al. 2008). Under wetter (or less arid) conditions we might expect i) an
690 expanded riparian fringe; ii) a greater catchment representation of C₃ plants; and/or iii) alleviated
691 water stress of this vegetation. Such a scenario would account for the significantly lower $\delta^{13}\text{C}$
692 values evident in the shallower sediment (P1 < P2 < P3; total variance of ~2 ‰; Fig. 5). Similar 2–
693 3 ‰ magnitude shifts in the $\delta^{13}\text{C}$ values of long (C_{29–33}) odd chain *n*-alkanes from Lake Malawi,
694 equatorial East Africa, was reported to reflect a 15 % increase in C₄ vegetation and millennial
695 drought during the Younger Dryas cold period (Castañeda et al. 2007), and in the $\delta^{13}\text{C}$ of long
696 chain *n*-alkanoic acids extracted from marine sediment was associated to a ~20 % variability in the
697 C₃ and C₄ vegetation of northeast Africa (Feakins et al. 2007). The higher relative abundances of
698 lower MW *n*-alkanes are also indicative of aquatic sources (Fig. 3). Consequently, sediment
699 biomarkers and *n*-alkanes $\delta^{13}\text{C}$ from the 14 Mile Pool effectively reflected the local response of
700 vegetation to hydrological and regional hydroclimatic change in the late Holocene, even though
701 this change may have been relatively modest on a catchment scale. Take together, our results
702 show that improved baselines and a careful consideration of the complex spatial and temporal
703 interaction between plants and water in arid tropical ecosystems are crucial for an appropriate
704 interpretation of past vegetation dynamics from molecular proxies.

705

706 **5. CONCLUSIONS**

707 The major organic input to sediments from an episodic wetland (Fortescue Marsh, Pilbara) (i.e.,
708 aquatic microflora, C₄ grasses, and C₃ plants originating at least partly from a significant riparian
709 fringe) could be revealed from the trace levels of aliphatic hydrocarbon compounds (including *n*-
710 alkanes, branched alkanes, hopanoids and steroids) detected in them. A temporal flux of some
711 sources was also evident. The stable isotopic signatures measured for several of the more

712 abundant *n*-alkanes provided additional C source information that can be useful for paleoclimate
713 reconstructions. Particularly, the relatively lower $\delta^{13}\text{C}$ values of long chain *n*-alkanes in modern
714 sediments reflected the emergence of wetter conditions through the late Holocene, and
715 particularly over recent decades. Though we highlight limitations and caution in interpreting
716 organic geochemical signals in sediments from tropical drylands, this approach provides valuable
717 information for the interpretation of arid paleoenvironmental records globally. Of note, these data
718 can serve as relatively modern analogues for the characteristics of biomarker records that have
719 been deposited during ancient dry phases recorded in very long (>100,000 yrs)
720 paleoenvironmental records, such as the ongoing projects of the International Continental
721 Scientific Drilling Program (<http://www.icdp-online.org/home/>): the Hominin Sites and Paleolakes
722 Drilling Project (HSPDP) in East Africa, the Lake Chad Deep Drilling (CHADRILL) in North Africa, the
723 Lake Petén Itza in Guatemala and the Lake Chalco in Mexico. Regionally, more extensive and
724 representative organic geochemical datasets, particularly of *n*-alkane $\delta^{13}\text{C}$ signatures from a range
725 of sources (e.g. riparian vegetation) common across northern and inland Australia and other sub-
726 tropical drylands, will help improve the accuracy and uncertainty of mixing models to reconstruct
727 past vegetation contributions quantitatively in these relatively underrepresented regions.

728

729

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- 1134

1135 **SUPPLEMENTARY MATERIAL**

1136

1137 **Table S1:** Carbon stable isotopic distributions of *n*-alkanes (*n*-C) for sediments from P1– P3 of the

1138 14 Mile Pool record

1139

<i>n</i> -C number	P1		P2		P3	
	$\delta^{13}\text{C}$	Std	$\delta^{13}\text{C}$	Std	$\delta^{13}\text{C}$	Std
17	-32.3	0.1	-	-	-	-
18	-	-	-	-	-	-
19	-31.3	1.1	-	-	-29.5	0.1
20	-30.7	0.2	-27.9	0.4	-30.0	0.4
21	-30.9	0.3	-28.4	0.5	-30.1	0.9
22	-29.8	0.2	-27.2	0.3	-30.0	0.7
23	-29.5	0.2	-27.8	0.1	-29.5	0.1
24	-30.1	0.1	-28.4	0.3	-30.1	0.3
25	-29.6	0.0	-28.1	0.2	-29.5	0.1
26	-29.7	0.1	-29.4	0.1	-29.7	0.2
27	-29.3	0.0	-28.8	0.05	-28.4	0.2
28	-29.6	0.1	-29.7	0.3	-29.7	0.3
29	-29.8	0.1	-29.0	0.1	-28.6	0.1
30	-29.5	0.2	-29.3	0.1	-29.5	0.1
31	-30.2	0.1	-29.5	0.1	-28.4	0.1
32	-30.7	0.1	-29.6	0.3	-29.9	0.2
33	-29.7	0.1	-28.9	0.1	-27.5	0.2
34	-29.7	0.4	-	-	-27.8	0.3
35	-26.6	0.3	-25.8	0.1	-24.0	0.1

1140

1141 (1-COLUMN FITTING TABLE)

1142 **Table S2:** Compilation of foliar $\delta^{13}\text{C}$ values (Fig. 6) from C_3 woody plants collected from the non-
 1143 riparian and riparian zone by the Ecosystems Research Group at The University of Western
 1144 Australia over the 1997–2013 period in the semi-arid Pilbara region.

1145

Species	Location	Collection date	n	$\delta^{13}\text{C}$ [‰VPDB]	Std
<u>Non-Riparian</u>					
<i>Acacia aneura</i>	Red Hill	01-11-1997	2	-25.6	0.49
<i>A. aneura</i>	Floodplain	01-11-1997	2	-25.5	0.11
<i>A. aneura</i> sens. lat.	Flat Rocks, Yandicoogina	17-23-05-2004	12	-26.9	0.77
<i>A. aneura</i> sens. lat.	Paraburdoo	17-23-05-2004	24	-26.4	1.23
<i>A. aneura</i> sens. lat.	West Angelas	01-08-2007	12	-27.5	0.48
<i>A. aneura</i> sens. lat.	West Angelas	01-08-2007	5	-27.4	0.42
<i>A. aneura</i> sens. lat.	West Angelas	01-08-2007	4	-27.5	1.01
<i>A. aneura</i> sens. lat.	West Angelas	01-08-2007	10	-28.2	0.89
<i>A. aneura</i> sens. lat.	West Angelas	01-08-2007	13	-27.6	0.91
<i>A. aneura</i> sens. lat.	West Angelas	01-08-2007	13	-28.2	0.64
<i>A. catenulata</i>	West Angelas	01-08-2007	1	-28.9	-
<i>A. catenulata</i>	West Angelas	01-08-2007	11	-28.8	0.62
<i>A. catenulata</i>	West Angelas	01-08-2007	6	-27.8	0.47
<i>A. catenulata</i>	West Angelas	01-08-2007	6	-27.4	0.60
<i>A. citrinoviridis</i>	Barnett Ck	01-11-1997	2	-26.9	0.85
<i>A. xiphophylla</i>	Floodplain	01-11-1997	2	-26.8	0.53
<i>A. xiphophylla</i>	Red Hill	01-11-1997	2	-26.2	0.04
<i>Callitris columellaris</i>	Callitris Juna Downs	01-05-2003	61	-26.9	0.82
<i>C. columellaris</i>	Callitris Juna Downs	01-05-2004	108	-27.3	0.76
<u>Riparian</u>					
<i>Eucalyptus camaldulensis</i>	Marillana Ck Snooker	17-04-2011	5	-29.9	0.98
<i>E. camaldulensis</i>	Yule River control	11-11-2010	3	-31.2	0.57
<i>E. leucophloia</i>	Red Hill	01-11-1997	2	-25.5	0.27
<i>E. victrix</i>	Barnett Ck	01-11-1997	2	-28.7	0.20
<i>E. victrix</i>	Marillana Ck Snooker/control	17-04-2011	5	-30.6	0.90
<i>E. camaldulensis/E. victrix</i>	Yule River control	28-04-2011	5	-31.5	0.44
<i>E. camaldulensis/E. victrix</i>	Weeli Wolli Ck	02-02-2012	24	-30.9	1.24
<i>E. camaldulensis/E. victrix</i>	Weeli Wolli Ck	16-05-2012	21	-31.9	0.93
<i>E. camaldulensis/E. victrix</i>	Weeli Wolli Ck	21-08-2012	10	-31.8	1.09
<i>E. camaldulensis/E. victrix</i>	Weeli Wolli Ck	16-11-2012	20	-31.6	0.99
<i>E. camaldulensis/E. victrix</i>	Weeli Wolli Ck	27-03-2013	47	-31.8	0.98
<i>Eucalyptus</i> spp.	Ratty Springs, Pirraburdoo Ck	24-11-2005	9	-27.4	0.24
<i>Melaleuca argentea</i>	Marillana Ck	15-04-2001	20	-30.2	1.37
<i>M. argentea</i>	Marillana Ck Snooker/control	29-11-2009	5	-31.6	0.52
<i>M. argentea</i>	Marillana Ck Snooker/control	24-07-2010	5	-32.0	0.55
<i>M. argentea</i>	Marillana Ck Snooker/control	17-04-2011	5	-31.6	0.78
<i>M. argentea</i>	Weeli Wolli Ck	15-04-2001	20	-30.7	0.43
<i>M. argentea</i>	Yule River control	01-03-2009	5	-32.1	0.85
<i>M. argentea</i>	Yule River control	01-06-2009	5	-33.5	0.31
<i>M. argentea</i>	Yule River control	01-11-2009	5	-32.9	0.69
<i>M. argentea</i>	Yule River control	13-05-2010	5	-33.1	0.76
<i>M. argentea</i>	Yule River control	11-11-2010	3	-32.7	0.55
<i>M. argentea</i>	Yule River control	28-04-2011	5	-33.4	0.59
<i>M. argentea</i>	Marillana Creek	2001	10	-30.3	1.05
<i>M. argentea</i>	Weeli Wolli Ck	2001	10	-30.7	0.32

1146

1147 (1.5-COLUMN FITTING TABLE)